Atmospheric CH₄ oxidation by Arctic permafrost and mineral cryosols as a function of water saturation and temperature

B. Stackhouse¹; M.C.Y. Lau¹; T. Vishnivetskaya², N. Burton; R. Wang; A. Southworth; L. Whyte³; T.C. Onstott¹

¹Department of Geosciences, Princeton University, NJ, USA

²The Center for Environmental Biotechnology, University of Tennessee, Knoxville, TN, USA

³Department of Natural Resource Sciences, McGill University, QC, Canada

N. Burton, R. Wang, and A. Southworth are not currently affiliated with research institutions but were at Princeton during the time of the research.

Corresponding author: T.C. Onstott

Mailing address: B79 Guyot Hall, Dept. of Geosciences, Princeton University, Washington Road, Princeton, NJ 08544

Email address: tullis@princeton.edu

Telephone: 609-258-7678

Fax: N/A

Autho

This is the author manuscript accepted for publication and has undergone full peer review but has not been through the copyediting, typesetting, pagination and proofreading process, which may lead to differences between this version and the <u>Version of Record</u>. Please cite this article as <u>doi:</u> 10.1111/gbi.12193

1	
2	Article Type: Original Article
3	Atmospheric CH ₄ oxidation by Arctic permafrost and mineral cryosols as a function of
4	water saturation and temperature
5	Running Title: Atmospheric CH ₄ oxidation by Arctic mineral cryosols
6	
7	B. Stackhouse ¹ ; M.C.Y. Lau ¹ ; T. Vishnivetskaya ² , N. Burton ¹ ; R. Wang ¹ ; A. Southworth ¹ ; L.
8	Whyte ³ ; T.C. Onstott ¹
9	
10	¹ Department of Geosciences, Princeton University, NJ, USA
11	² The Center for Environmental Biotechnology, University of Tennessee, Knoxville, TN, USA
12	³ Department of Natural Resource Sciences, McGill University, QC, Canada
13	
14	Keywords: methanotrophs, stable isotope labeling, microcosms
15	
16	
17	Abstract (200 words)
18	The response of methanotrophic bacteria capable of oxidizing atmospheric CH ₄ to climate
19	warming is poorly understood, especially for those present in Arctic mineral cryosols. The
20	atmospheric CH ₄ oxidation rates were measured in microcosms incubated at 4°C and 10°C along
21	a 1 m depth profile and over a range of water saturation conditions for mineral cryosols

- 22 containing Type I and Type II methanotrophs from Axel Heiberg Island (AHI), Nunavut,
- 23 Canada. The cryosols exhibited net consumption of \sim 2 ppmv CH₄ under all conditions, including
- 24 during anaerobic incubations. Methane oxidation rates increased with temperature and decreased

with increasing water saturation and depth, exhibiting the highest rates at 10°C and 33%

saturation at 5 cm depth (260 \pm 60 pmol CH₄ gdw⁻¹ d⁻¹). Extrapolation of the CH₄ oxidation rates

to the field yields net CH₄ uptake fluxes ranging from 11 to 73 μ mol CH₄ m⁻² d⁻¹, which are

comparable to field measurements. Stable isotope mass balance indicates ~50% of the oxidized

29 CH₄ is incorporated into the biomass regardless of temperature or saturation. Future atmospheric

30 CH₄ uptake rates at AHI with increasing temperatures will be determined by the interplay of

31 increasing CH₄ oxidation rates versus water saturation and the depth to the water table during

32 summer thaw.

33 Introduction

The response of terrestrial ecosystems, especially those located in the Arctic, to rapid 34 global warming is the subject of much debate that centers on the impact of future greenhouse gas 35 emissions and feedbacks (Lawrence et al., 2008; Allison et al., 2010; Graham et al., 2012). 36 37 Arctic permafrost and overlying active layers contain approximately one third of the global soil organic carbon (SOC) stocks shallower than 3 meters (Schuur et al., 2008). These reserves will 38 39 be particularly impacted by climate change as deepening of the seasonally thawed active layer reintroduces the previously sequestered permafrost SOC to the active carbon cycle (Rowlands et 40 al., 2012). Based upon modeling experiments, Hayes et al. (2014) have estimated that Arctic 41 warming from 1970 to 2006 has already exposed an additional 11.6 Pg of SOC to microbial 42 43 degradation, and that this exposure has resulted in the net emission of 3.7 Pg of CO₂ and 0.03 Pg 44 of CH₄. Not considered in their model are the impacts of slumps associated with permafrost degradation that create new hydrological conditions, such as thermokarst lakes or dry upland 45 cryosols, throughout Arctic regions (Jorgenson et al., 2006; Osterkamp, 2007; Grosse et al., 46 2011). The expansion of thermokarst lakes and flooded cryosols in Siberian permafrost where 47 48 Pleistocene-age SOC is undergoing rapid conversion to CH₄ under anaerobic conditions has been tied to increases in regional CH₄ emissions (Walter et al., 2006). Although atmospheric CH₄ 49 concentrations at a global scale remained relatively stable at 1.78 ppmv from 1999 to 2006, they 50 have increased from 2007 to 2010 to 1.8 ppmv (Dlugokencky et al., 2014). CH₄ has a global 51 warming potential 75 times greater than that of CO₂ over a 20-year timescale (Shindell et al., 52 2009) and at 1.8 ppmv, atmospheric CH₄ accounts for ~18% of the greenhouse gas radiative 53 forcing (Montzka et al., 2011). This recent upward trend in atmospheric CH₄ concentrations 54

55 appears to be caused by global meteorological factors and is not due to release from Arctic SOC stocks even though Arctic temperatures have been steadily increasing during this time period 56 (Dlugokencky et al., 2014). This apparent paradox may stem from the fact that the majority of 57 field research on Arctic CH₄ fluxes has been performed at sites in Siberia, Greenland, and Alaska 58 where a high-SOC, CH₄-emitting active layer overlies permafrost that is slowly thawing (Table 59 1). Various other Arctic sites, however, have recently been found to be net sinks of atmospheric 60 CH₄ (Whalen & Reeburgh, 1990, 1992; Adachi et al., 2006; Kolb, 2009; Bäckstrand et al., 2010; 61 Liebner et al., 2011; Brummell et al., 2012, 2014; Allan et al., 2014; Emmerton et al., 2014; Lau 62 et al., 2015; Stackhouse et al., 2015, See Table 1). With a few exceptions these sites are located 63 in mineral cryosols that contain $<50 \text{ kg SOC m}^{-2}$ and which comprise 87% of permafrost-64 impacted regions (Hugelius et al., 2014). The ability of soils to oxidize CH₄ from the atmosphere 65 has also been increasingly connected to the presence of high-affinity methanotrophic bacteria 66 based upon detection of a specific pmoA marker gene (e.g. Upland Soil Cluster alpha) (Knief & 67 Dunfield, 2005; Kolb, 2009; Degelmann et al., 2010; Martineau et al., 2014; Lau et al., 2015), 68 even though these bacteria have not yet been isolated in the lab. Soils containing high-affinity 69 methanotrophic bacteria exhibit K_m values of 10-280 nM CH₄ (Knief & Dunfield, 2005), low 70 V_{max} of ~0.1 fmol CH₄ cell⁻¹ h⁻¹ (Baani & Liesack, 2008) and threshold CH₄ mixing ratios for 71 CH₄ oxidation of ~0.1 ppmv (Bender & Conrad, 1992; Whalen & Reeburgh, 1992; Roslev et al., 72 1997). The reason for this threshold in CH₄ concentration is still not understood since at this 73 74 concentration the methanotroph energy yield rate is greater than that estimated for cell 75 maintenance (Kolb et al., 2005). The threshold CH_4 mixing ratios for cell growth, however, have 76 been determined to range from 10 to 100 ppmv for two Methylocytis strains (Knief & Dunfield, 2005). One possible explanation for the apparent CH₄ oxidation threshold observed in some soil 77 78 studies is that it may reflect a dynamic balance between CH₄ oxidation by methanotrophs and CH₄ production by methanogens as revealed by dual isotope studies (Ambus et al., 2002). 79

The models used by the Intergovernmental Panel on Climate Change (IPCC AR4) did not account for increases in available SOC due to permafrost thawing or for feedbacks related to permafrost CH₄ emissions (Koven et al., 2011). Models that do take these into account predict greater CH₄ emissions and that the Arctic will become a net source of CO₂ rather than a sink of CO₂ by 2100 (Koven et al., 2011). But even these models may be biased because they have not taken into consideration the results from mineral cryosols sites where atmospheric CH₄ oxidation

has been reported. We, therefore, used a combination of microcosms, stable isotope mass 86 balance, and metagenomic information, to measure the atmospheric CH_4 oxidation rates of a 87 mineral cryosol from the Canadian high Arctic where atmospheric CH_4 uptake has been reported 88 in the field and in the lab (Allan et al., 2014; Stackhouse et al., 2015) and where high-affinity 89 methanotrophic bacteria have been detected (Martineau et al., 2014; Lau et al., 2015). We sought 90 to examine the effects of temperature, water saturation, O₂ potential, light exposure, acetate 91 addition, bicarbonate addition, and depth on the CH₄ oxidation rates and to assess the growth 92 yield of atmospheric CH₄ oxidizers. 93

94

95 Materials and Methods

96 Field Site

Samples were collected in April 2011 prior to spring thaw from an upland ice-wedge polygon at 97 98 the McGill Arctic Research Station (MARS) located on Axel Heiberg Island (AHI), Nunavut, Canada (N79°24', W90°45'). Cryosols in this area are classified as acidic (pH 5 – 6.5), static and 99 100 turbic cryosols and have a sparse coverage of lichen, Salix artica, Polygonum viviparum, and the genera Dryas, Saxifraga, Papaver, and Eriophorum (Stackhouse et al., 2015). The mean annual 101 air temperature (MAAT) is -17.8°C (Pollard, personal communication, 2010). This is one degree 102 103 warmer than the -18.8°C annual mean air temperature for 2010 of the Eureka weather station 104 approximately one half degree latitude further north (Canadian Climate Normals 1981–2010. Environment Canada. Climate ID: 2401200). During July 2012 air temperatures were observed 105 to reach 15°C and cryosol temperatures ranged from 0°C at the permafrost table to 8°C at the 106 surface. Active layer depth at this location from 2009-2012 has varied between 50-70 cm during 107 maximum active layer development in July. Cryosol SOC was found to vary between 1-6% in 108 109 the top 10 cm and was <1.5% from 11 cm down to 100 cm depth with a C:N ratio that ranged from 16 to 70 in the top 10 cm and from 15 to 19 from 11 cm to 100 cm (Stackhouse et al., 110 2015). The δ^{13} C of the SOC was -26.5±0.5‰ and exhibited no trend with depth (Stackhouse et 111 al., 2015). Undegraded root, stem, and leaf material was observed in the top 10 cm of the 112 cryosol. The 1 m cores were collected into 70% EtOH-sterilized 3" polycarbonate tubes inside of 113

- a SIPRE coring barrel (Jon's Machine Shop, Fairbanks, Alaska), capped, and transported frozen
 back to Princeton University where they were stored at -20° C until use.
- 116

117 Microcosm preparation

In a 4.5°C cold room a representative frozen core from the polygon interior was partitioned 118 into 10 cm sections using a miter saw sterilized with 70% EtOH. Frozen sections of the core 119 from 0-10 cm (highest SOC content), 30-40 cm (midway point between surface and permafrost), 120 60-70 (permafrost table), and 70-80 cm (permafrost depth) were chosen for creating microcosms. 121 Core subsections were allowed to thaw on ice for 20 minutes before the outer layer of core 122 sections that had been in contact with the polycarbonate tubing was removed using a sterile 123 knife. Remaining cryosol was placed into Whirl-Pak® bags and placed on ice until thawed. The 124 10 cm sections were then homogenized by mixing in the Whirl-Pak® bags to avoid subsampling 125 heterogeneous portions of the soil. 126

The 160 mL serum vials used for the microcosm experiments were pretreated by soaking in 127 128 10% HNO₃ overnight, triple-rinsing with distilled H₂O, and then baking at 450°C for 8 hours to remove any organic carbon contamination. To prevent degassing of butyl rubber stoppers into 129 130 the vials, new stoppers were boiled in 0.1 N NaOH for 1 hour, triple rinsed with distilled H_2O_1 , autoclaved, and then soaked overnight in distilled H₂O following the protocol of Oremland et al. 131 (1987). In a 4.5°C cold room and on ice, 8-10 g of cryosol was placed in the serum vials and 132 capped with stoppers and crimp sealed. Vials were then flushed at 20 PSI for 3 minutes and 133 pressurized to 1.5 atm with a gas mixture containing 400 ppmv CO₂, 1.8 ppmv CH₄, 20% O₂, 134 80% N₂ (AirGas) for aerobic incubation and with UHP N₂ for anaerobic incubation. Serum vials 135 were covered in aluminum foil to prevent photosynthetic fixation and then incubated in triplicate. 136

The water content of the mineral cryosol was determined by gravimetric difference after drying the homogenized samples in triplicate at 80°C overnight. Initial water contents of the mineral cryosol were found to be $30.4 \pm 2.0\%$ (0-10 cm), $11.0 \pm 0.4\%$ (30-40 cm), $8.4 \pm 0.6\%$ (60-70 cm), $20.4 \pm 0.7\%$ (70-80 cm). Because the highest water holding content (i.e. water saturation) of the cryosol was ~33% H₂O w/w, microcosm experiments were prepared with 33%, 66%, and 100% of water saturation (equating to ~11%, 22%, and 33% H₂O w/w, respectively)

143 for all depths and incubated in triplicate at 4°C and 10°C in temperature-monitored refrigerators. Experiments that were run with cryosols at below in situ water saturation had cryosol samples 144 dried in a desiccator at 4.5°C without stirring until reaching the desired water content as 145 determined by gravimetric difference. Desiccation required up to 8 days. Although the transition 146 from freezing to 4.5°C during this time may allow for shifts in the microbial community before 147 the experiment began, this was judged to be the least disruptive manner to dry the soils before 148 beginning the microcosm incubations. Cryosols with in situ water content below 66% and 100% 149 150 saturation level were adjusted to that level using DI H₂O. Treatments were run for 35 (30-40 cm, 70-80 cm), 31 (0-10 cm), 20 (60-70 cm) days. 151

Anaerobic 70-80 cm cryosol samples were run for 52 days under a headspace of UHP N_2 with the headspace adjusted to either 2.6 ppmv CH₄ or 10 ppmv CH₄ at in situ water saturation (~100%) at both 4°C and 10°C in temperature-monitored refrigerators. Aerobic and anaerobic microcosms were run with vials containing no cryosol additions as controls for outgassing of stoppers, dilution effect due to multiple sampling, and to monitor instrumental drift and potential leakage. The CO₂ and CH₄ concentrations in vials with cryosol were corrected to match the observed change in the controls vials to isolate the microbial signal.

To evaluate the effect of substrate addition and light exposure on CO₂ and CH₄ emissions, 159 160 three additional sets of microcosm were prepared, using active layer cryosol exposed to light (AL-Light), active layer cryosol in the dark (AL-Dark), and permafrost in the dark (PF). The 161 microcosms contained 5±0.1 g of active layer (AL, 0-10 cm) or permafrost (PF, 70-80 cm) 162 samples. AL-Light and AL-Dark microcosms were incubated under 1.5 atm of ultra-zero grade 163 164 air (20% O_2 , 80% N_2 ; AirGas, Inc.) containing <10 ppmv CO₂, <50 ppbv CH₄, <30 ppbv CO, and <30 ppbv H₂. After 6 days of incubation, an 8 ppmv CH₄ spike was added to the headspace 165 166 of the AL-Light and AL-Dark microcosm sets to test for CH₄ oxidation potential. PF microcosms were incubated under 1.5 atm of ultrahigh purity (UHP) N₂ (AirGas Inc.) to test for the 167 methanogenic potential of thawing permafrost. Each microcosm set contained four treatments, 168 without any amendments (in situ), with 2.5 mL of MilliQ water (exceeding the water saturation 169 170 of the cryosol), 2.5 mL of 1 mM acetate solution (Ac), or 2.5 mL of 1 mM NaHCO₃ 171 (Bicarbonate), replicated in duplicate. Preparation and incubation was carried out in a 4.5°C cold room. Vials were flushed again after the addition of amendments and pressurized to 1.5 atm. 172

- 173 Two vials containing no cryosols (blank) were also included as control. All vials were wrapped
- in aluminum foil and incubated in complete darkness (AL-dark and PF). AL-Light microcosms
- were incubated under 24 h illumination (Photosynthetic available radiation, $PAR = 37 \mu mol m^{-2}$
- 176 s^{-1}). These microcosms were run for 42 days.
- 177

178 CO₂ and CH₄ Measurements

- CO₂ and CH₄ concentrations of the aerobic microcosms were measured by iCO₂-CRDS (cavity 179 ring-down spectrometer, Picarro G2101-i CRDS, USA) using a small sample inlet module 180 (SSIM). Measurements were made on injections of 15 mL of gas and headspace was replaced 181 with the original gas mixture to maintain a 1.5 atm pressure inside the vials. Sample injection 182 into the SSIM with the G2101-i coordinator dilutes the sample with 5 mL of ultra-zero grade air 183 (AirGas, Inc.) to bring sample volume up to 20 mL during measurements. Reported 184 concentrations were corrected to account for this dilution. The CH₄ concentrations for anaerobic 185 microcosms were measured by flame ionization (FID) (detection limit ~100 ppbv, FID Peak 186 Performer 1 series, Peak Laboratories, USA) using UHP Argon as a carrier gas, and the CO₂ was 187 measured using a methanizer and FID on the same instrument. 188
- 189

190 Stable Isotope Experiments

- 191 After the initial incubation experiments, those microcosms from 0-10 cm depth exhibiting
- atmospheric CH₄ oxidation were flushed and pressurized to 1.5 atm with ultra-zero grade air.
- 193 The vials were then spiked to a concentration of ~10 ppmv with 99% atom labeled ${}^{13}CH_4$ (${}^{13}C =$
- 194 8.8×10^9 ‰ VPDB) and were incubated once more at the original temperatures of 4°C and 10°C.
- 195 The δ^{13} C-CO₂ and the CO₂ concentrations were measured by CRDS (Picarro G2101-i CRDS,
- 196 USA) using 20 mL STP of the headspace gas. Isotopic tracer calculations of CH₄ oxidation rates
- 197 were performed by mass balance, assuming that 1) the δ^{13} C of CO₂ produced by heterotrophic
- activity was equal to that of the cryosol SOC δ^{13} C (-26.5±0.5‰ VPDB) and 2) that all additional
- 199 13 C-CO₂ was derived from the oxidation of 13 CH₄. The 13 C/ 12 C ratio of headspace CO₂ was
- 200 calculated by:

$$\frac{{}^{13}C}{{}^{12}C}_{CO2} = \left(\frac{\partial {}^{13}C_{CO2}}{1000} + 1\right) * \frac{{}^{13}C}{{}^{12}C}_{VPDB}$$
(1)

where the ${}^{13}C/{}^{12}C_{VPDB}$ ratio is 0.0112372. From this the total moles of ${}^{13}C$ in the headspace can be calculated by:_

$$moles {}^{13}C_{Total} = \frac{moles CO2_{Total} * \frac{{}^{13}C}{{}^{12}C_{CO2}}}{\left(1 + \frac{{}^{13}C}{{}^{12}C_{CO2}}\right)}$$
(2)

203 The moles of 13 C in the headspace derived from oxidation of the SOC are calculated as:

moles
$${}^{13}C_{From \, soil} = \frac{moles \, CO2_{Total} * \frac{{}^{13}C}{{}^{12}C_{Soil}}}{\left(1 + \frac{{}^{13}C}{{}^{12}C_{Soil}}\right)}$$
(3)

204

The difference in the total ¹³C from the ¹³C expected from SOC degradation is assumed to be the product of CH₄ oxidation to CO₂ by the methanotrophic microbial population since all ¹²C added to the headspace is assumed to be derived from SOC:

$$moles \ {}^{13}C_{Total} - moles \ {}^{13}C_{From \ soil} = moles \ {}^{13}C_{CH4 \ oxidation} \tag{4}$$

This mole abundance was used to determine a rate of CH₄ oxidation independent of the direct measurements of CH₄ consumption in the headspace. CH₄ loss was directly measured by FID using UHP Argon as a carrier gas. CH₄ oxidation rates calculated from the measured headspace CH₄ concentrations were then compared to the rates determined by the shift in δ^{13} C enrichment of headspace CO₂. The difference between the amount of ¹³CH₄ consumed and the amount of ¹³CO₂ generated was used to estimate the amount of ¹³C incorporated into the biomass, assuming that a minor fraction of respired ¹³CH₄ was lost to the system as dissolved organic carbon.

215

216 Rate Constant Calculations

First order rate constants were calculated for CH₄ oxidation at 4°C and 10°C at 33%, 66%, and 100% water saturation. Constants were determined from the slope of ln [CH₄] versus time

and normalized to cryosol mass, resulting in units of $s^{-1} g^{-1}$. The slope was calculated from the

- 220 linear period of active CH₄ oxidation, avoiding any initial lag phase and final threshold level in
- 221 oxidation. First order reaction rates were also derived by simulation of the results using the
- 222 Geochemist's Workbench V8.0.12 (Bethke, 2008). Microcosm simulations were carried out for
- the aerobic 0-10 cm cryosol microcosms under all saturation conditions and for the anaerobic
- 224 permafrost microcosms. CH₄ consumption was modeled using the following reaction:

$$CH_4 + 2O_2 \rightarrow HCO_3^- + H^+ + H_2O$$

and the following dual-Monod model developed by Jin and Bethke [2003],

$$\mathbf{k} = \mathbf{V}_{\max} \left[\mathbf{X} \right] \mathbf{F}_{\mathrm{D}} \mathbf{F}_{\mathrm{A}} \mathbf{F}_{\mathrm{T}} \tag{5}$$

where the reaction rate, k is in mol (g of soil)⁻¹ s⁻¹, V_{max} is the specific cellular rate for the CH₄ oxidization reaction in mol (mg of biomass)⁻¹ s⁻¹, [X] is the methanotroph concentration in mg of biomass (g of soil)⁻¹ of the atmospheric CH₄ oxidizers. F_D is the parameter controlling the electron donating reaction rate and was defined by the equation

$$F_D = m_D / (m_D + K_D) \tag{6}$$

where m_D is the concentration of CH₄. Similarly, the electron-accepting reaction rate was controlled by the parameter, F_A , defined by

$$F_A = m_A / (m_A + K_A) \tag{7}$$

233

where m_A is the concentration of O_2 . The terms K_D and K_A represent the half-saturation Michaelis constants, K_m , reported in M units. A V_{max} of 310 nmol g-biomass⁻¹ s⁻¹ was chosen based on the V_{max} measured by Baani and Liesack (2008) for the high-affinity pMMO of Methylocystis spp. SC2 and assuming a cell dry weight mass of 10^{-13} g cell⁻¹ (Phelps et al., 1994). A K_A for O_2 of 0.14 μ M was chosen based upon measurements of Type II methanotroph strain OU-4-1 (Joergensen, 1985). A K_D of 0.11 μ M was derived from the measurements of Baani and Liesack (2008).

The thermodynamic potential factor, F_T , was used to model the threshold CH_4 concentration where CH_4 oxidation ceased and was defined as

$$F_{\rm T} = 1 - \exp[-f/(xRT)]$$
 (8)

243

where R is the real gas constant 8.314472 J K^{-1} mol⁻¹, T is the temperature in K, x is the average stoichiometric number (Temkin, 1963), or the ratio of the free energy change of the overall reaction to the sum of the free energy changes for each elementary step. For this study we assumed that x was 1 for the reaction stoichiometry as written above. f is the net thermodynamic driving force of the reaction defined by,

$$f = -\Delta G - m \Delta G_p \tag{9}$$

249

where ΔG is the free energy change of the redox reaction, m is the number of moles of ATP 250 generated per mole of reaction and ΔG_p is the free energy for the phosphorylation reaction ADP 251 + P \rightarrow ATP (Thauer et al., 1977). For aerobic CH₄ oxidation, ATP is generated by a trans-252 membrane H⁺ gradient created by an electron transport chain. The oxidation of one CH₄ shuttles 253 9-10 H^+ to the periplasm thereby generating ~3.3 ATP (Bürgmann 2011). The Gibbs free energy 254 for the formation of ATP at 4°C and at a pH of 5 is ~60 kJ (mole of ATP)⁻¹ (Larowe and 255 Helgeson, 2007). According to equation 9 this would correspond to a theoretical minimum free 256 energy threshold of ~193 kJ per mole of CH₄ oxidized. 257

258 The active biomass concentration was calculated using the following relationship,

259

d[X]/dt = Y k - [X] D (10)

where Y is the yield in mg of biomass mol⁻¹ of CH_4 oxidized, which was determined experimentally, k is the reaction rate from equation (5), and D is the death rate in s⁻¹, which in this case was assumed to be zero given the short duration of the experiments.

Simulated CH_4 fugacity was fit to experimental data by utilizing a reaction volume that matched the total CH_4 available in the microcosms, modifying the reaction rate, the initial active biomass, growth yield, ΔG_{ATP} of the reaction, and the thermodynamic potential factor (F_T) of the reaction, while gas concentrations were initialized based on the measured partial pressures. The F_T is threshold free energy below which the reaction ceases (Bethke, 2008). In order to account

268 for Geochemist's Workbench program not mass balancing the formation of biomass, newly created biomass C was subtracted from the carbon pool at each time step of the model. This 269 270 resulted in a small decrease of up to 50 ppbv CH₄ from the headspace but did not substantially alter the modeled results. First order reaction constants (k) were calculated from the modeled 271 reaction by dividing the maximum rate of CH₄ oxidation in each microcosm by the dissolved 272 CH_4 concentration at that point and normalized to cryosol mass, resulting in units of s⁻¹ g⁻¹. The 273 activation energy for CH₄ oxidation was calculated by multiplying the ideal gas constant (8.314 J 274 K^{-1} mol⁻¹) by the slope of plots of ln(k) against 1000/T(K). 275

The addition of acetoclastic methanogenesis into the CH₄ oxidation model as an
additional source term was tested using the following reaction

Acetate +
$$H_2O \rightarrow CH_4(aq) + HCO_3^-$$
 (11)

using the parameters of Methanothrix, a K_D of 86 μ M (Jetten et al., 1990) and an acetate

threshold derived from Westermann et al. (1989).

The CO_2 emission rates were calculated by simply dividing the final CO_2 emitted in moles by the total duration of incubation and the dry weight of the cryosol.

282 Temperature Response Calculation

The temperature coefficient, Q_{10} , was calculated for the 0-10 cm microcosm incubations by the following formula

$$Q_{10} = \left(\frac{R_2}{R_1}\right)^{\left(\frac{10}{T_2 - T_1}\right)}$$
(52)

where R_1 and R_2 stand for the rate of the reaction at the temperatures T_1 and T_2 , respectively. Similar to the rate constant calculations, R_1 and R_2 were selected during the period of active oxidation for determining CH₄ oxidation rates.

288

289 Cell Count

- 290 Cryosol cellular abundances were measured by fluorescent in situ hybridization (FISH). Soil
- samples were analyzed from frozen cryosol at 5 cm, 35 cm, 65 cm, and ~80 cm with 4 replicates

- from each, for a total of 16 samples. For FISH analysis, the following probes were used for
- eukaryotic cells (EUK516 tagged with Alexa-633 [5'-ACCAGACTTGCCCTCC-3'] targeting
- eukaryal 18S rRNA genes), archaeal cells (ARC915 tagged with Alexa-555 [5'-
- 295 GTGCTCCCCGCCA ATTCCT-3'] targeting archaeal 16S rRNA genes), and bacterial cells
- 296 (EUB338 tagged with Alexa-594 [5'-GCTGCCTCCCGTAGGAGT- 3'] targeting bacterial 16S
- 297 rRNA genes). Sample preparation and FISH staining was performed as described
- 298 (Vishnivetskaya et al., 2014).
- 299
- 300 **Results**
- 301 CH₄ oxidation

Aerobic incubations - The aerobic CH₄ oxidation rates for the 0-10 cm samples ranged 302 from 16 ± 12 (k=0.003±0.002 g⁻¹ d⁻¹) to 260 ± 60 (k=0.036±0.005 g⁻¹ d⁻¹) pmol CH₄ gdw⁻¹ d⁻¹, with 303 304 the 10°C and 33% saturation treatment having the highest observed rate of any treatments (ANOVA, p=0.001, Table 2, Fig. 1a). The 10°C and 33% saturation treatment for the 30-40 cm 305 samples also yielded the highest CH_4 oxidation rate for that depth (ANOVA, p=0.005) at 42±3 306 pmol CH₄ gdw⁻¹ d⁻¹ (Fig. 1b). The enhancement of the aerobic CH₄ oxidation rate for the 33% 307 saturation treatment was not observed for the 60-70 cm and 70-80 cm samples, nor was it 308 observed for any of the 4°C treatments. Although samples from lower depths exhibited low or no 309 CH₄ oxidation, none of the aerobic incubations produced CH₄ with statistical certainty (Table 2). 310

The 4°C microcosms showed a consistent delay of ~7 days before CH_4 oxidation began, whereas the 10°C, 33% saturation microcosms exhibited CH_4 oxidation by the third day of incubation and completed CH_4 drawdown within 10 days. CH_4 oxidation ceased at threshold concentrations between 0.1 and 0.5 ppmv CH_4 in the 10°C microcosms and between 0.5 and 1.4 ppmv CH_4 in the 4°C microcosms. The temperature response (Q_{10}) of the 0-10 cm CH_4 oxidation rate was 24.2±12.3 in the 33% saturation samples, 8.3±6.9 in the 66% saturation samples, and 2.4±1.6 in the 100% saturation samples.

For 0-10 cm and 30-40 cm, water saturation and temperature were found to have
significant effects on CH₄ oxidation with a significant interaction between the variables such that

320 the oxidation rate increased with temperature and decreased with water saturation (two-way

ANOVA, p <0.001, Table 3). Microcosms from 60-70 cm exhibited significant increases with

- temperature (p<0.001), but did not exhibit any significant effect from water saturation, and no
- interaction between the two variables was observed (Table 3). Neither temperature nor water
- saturation level had a significant effect on aerobic CH_4 oxidation in the 70-80 cm samples.
- Anaerobic incubations The anaerobic microcosms of the 70-80 cm cryosols with in situ water saturation levels yielded non-detectable anaerobic CH₄ oxidation (4°C and 10°C with 10 ppmv CH₄; Table 2) to -8.6 \pm 1.4 pmol CH₄ gdw⁻¹ d⁻¹ (4°C and 10°C with 1.8 ppmv CH₄; Table 2). CH₄ production above the concentration of the initial CH₄ spike was not observed in any anaerobic microcosms (Fig. S1g-h). For the 70-80 cm depth, water saturation and temperature were not found to have significant effects on CH₄ oxidation, nor was there a significant interaction between the variables (Table 3).
- Light exposure and substrate addition incubations- AL microcosms examining the effects 332 of light and substrate exposure exhibited clear CH₄ consumption. CH₄ oxidation rates were 333 highest in the in situ AL-Light and in situ AL-Dark microcosms (288 \pm 123 pmol CH₄ g⁻¹ d⁻¹ 334 $(k=0.015\pm0.004 \text{ g}^{-1} \text{ d}^{-1})$ and $411\pm70 \text{ pmol CH}_4 \text{ g}^{-1} \text{ d}^{-1}$ (k=0.022±0.002 g⁻¹ d⁻¹), respectively). 335 Final headspace concentrations of CH₄ ranged from 0.06±0.01 to 3.3±0.3 ppmv. CH₄ oxidation 336 occurred at a higher rate in AL-Dark microcosms for the acetate and bicarbonate treatments 337 (p=0.02 and 0.03, respectively) (Fig. 2a). By the end of the experimental run, AL microcosms 338 with amended soils yielded threshold concentrations CH₄ of 1.8-3.2 ppmv, compared to 0.05-0.9 339 ppmv in AL in situ microcosms. The light versus dark was not quite significant (p=0.06) though 340 amendment variables were found to be significant determinants of CH₄ oxidation (p=0.0004) and 341 342 no interactions between the variables were observed (p=0.4, Table 3).
- Anaerobic PF microcosms containing no CH_4 in the headspace were found to be neither sources nor sinks of CH_4 (Fig. 2a). The methanotrophic rates for the PF microcosms, with or without amendment, were not significantly different (Table 2; Fig. 2), with an average of 3 ± 10 pmol CH_4 g⁻¹ d⁻¹ averaging across all experimental conditions.
- ¹³CH₄ oxidation During the ¹³CH₄ incubation experiment, the ¹³CO₂ became enriched from $-16\pm6\%$ up to 400‰. The directly measured ¹³CH₄ oxidation rate of the 0-10 cm sample

349 was observed to be higher at 10°C than at 4°C (Table 2, p=0.05). No significant differences in the ${}^{13}CH_4$ oxidation rates were detected between water saturation levels in either the $4^{\circ}C$ 350 351 treatment (p = 0.55) or the 10^oC treatment (p = 0.21). When individual samples were examined, the ratio of expected CH₄ loss from isotopic mass balance of δ^{13} C of CO₂ to directly observed 352 CH₄ loss was ~0.5 (4°C, slope = $0.46 \pm 0.01 \text{ R}^2 = 0.99$; 10°C, slope = 0.56 ± 0.04 , R² = 0.97) 353 regardless of the total amount of CH₄ being oxidized, indicating that half of oxidized CH₄ 354 remained in the cryosol (Fig. 3a, 3b). Assuming that all of the oxidized CH₄ that did not become 355 356 CO_2 was incorporated into biomass, this would result in maximal growth yields of 7.4±0.2 g Cbiomass mol-CH₄⁻¹ at 4°C and 9.0 \pm 0.6 g C-biomass mol-CH₄⁻¹ at 10°C. The anaerobic ¹³CH₄ 357 experiment for the 70-80 cm sample yielded a CH₄ oxidation rate below the detection limit with 358 a shift in the δ^{13} C of the headspace CO₂ from -17±3‰ to 8±14‰. 359

360 CO₂ production

None of the aerobic microcosm experiments exhibited net CO_2 uptake even in the presence of light and all CO_2 emission rates exhibited a two-phase behavior, with a gradual decrease in the emission rates after ~10 days. The CO_2 production rates in the 0-10 cm cryosol samples were the highest among the measured depth profiles (p=0.05) and ranged from 98±17 to 235±150 nmol CO_2 gdw⁻¹ d⁻¹ with no difference observed between temperature and saturation treatments (p≥0.11) (Table 2, Table 3).

CO₂ production for 66% and 100% water saturation at 60-70 cm and 70-80 cm samples 367 ranged from 34 ± 2 to 44 ± 2 nmol CO₂ gdw⁻¹ d⁻¹. The 33% saturation samples at 4°C and 10°C for 368 the 30-40 cm, 60-70 cm, and 70-80 cm depths ranged from 13 ± 3 to 37 ± 5 nmol CO₂ gdw⁻¹ d⁻¹, 369 though all except for one treatment were below 23 nmol CO_2 gdw⁻¹ d⁻¹. These incubations 370 showed no difference as a function of temperature but water saturation level was found to 371 372 positively correlate with CO₂ emissions (Table 3). An interaction between the variables was found to exist for the 30-40 cm and 60-70 cm microcosms. The temperature response (Q_{10}) of the 373 374 CO_2 emission rate was 0.9±0.3 for the 33% saturation samples, 2.2±0.8 for the 66% saturation samples, and 3.2 ± 2.0 for the 100% saturation samples. 375

Anaerobic CO₂ production occurred linearly for the first 35 days at a rate of 36 ± 2 nmol CO₂ gdw⁻¹ d⁻¹ at 4°C and 47±14 nmol CO₂ gdw⁻¹ d⁻¹ at 10°C, after which time CO₂ emissions

reduced to near-zero rates until the end of the experiment at 52 days. During the aerobic ¹³CH₄ oxidation experiment, the 0-10 cm cryosol CO₂ production rates did not significantly vary by saturation level or by incubation temperature ($p\geq0.19$, Table 2). The anaerobic ¹³CH₄ oxidation experiment yielded CO₂ production rates of 19±5 nmol CO₂ gdw⁻¹ d⁻¹ at 4°C and 41±7 nmol CO₂ gdw⁻¹ d⁻¹ at 10°C, values in line with the emission rates of aerobic samples below 30 cm but less than those observed in 0-10 cm aerobic samples.

In microcosms examining the effects of light and substrate addition, CO₂ emission rates 384 in PF microcosms oscillated between instantaneous rates of <10 and 70 nmol CO₂ g⁻¹ d⁻¹. AL-385 light microcosms had a higher constant CO₂ emission rate (up to $420\pm63 \text{ nmol CO}_2 \text{ g}^{-1} \text{ d}^{-1}$) than 386 AL-dark microcosms (up to 179 ± 53 nmol CO₂ g⁻¹ d⁻¹), with the exception of AL-Light 387 bicarbonate additions (Table 2; Fig. 2b). AL soils produced at least five times more CO₂ than PF 388 soils (Table 2). Exposure to light significantly enhanced CO₂ emissions in the AL microcosms, 389 390 whereas amendments were not found to have a significant effect on AL or PF microcosms (Table 3). 391

392

393 Microcosm simulations

Data from the ¹³CH₄ microcosm experiments in this study indicated a maximum carbon use efficiency of ~50% (8 g-biomass C mol-CH₄⁻¹), which, assuming ~50% C cellular makeup by weight (Bratbak & Dundas, 1984) would result in a growth yield of ~16 grams of biomass per mole of oxidized CH₄. In our modeled CH₄ depletion curves we set the methanotrophic growth yield to 16 grams of biomass per mole of CH₄ oxidized for all conditions.

In fitting the modeled CH₄ oxidation curves to the experimental data the ΔG_{p} had to be 399 increased from 60 kJ (mole of ATP)⁻¹ to 285 to 287 kJ (mole of ATP)⁻¹ in order to recreate the 400 threshold CH₄ concentration observed for the 4°C and 33% water saturation microcosm (Fig. S1; 401 402 Table S1), otherwise the predicted CH₄ concentrations would decrease to sub-ppt levels. Alternatively models using a balance between methanotrophy and methanogenesis was tried in 403 404 order to create the observed threshold CH_4 concentrations (Fig. 4). Models incorporating the process of acetoclastic methanogenesis temporarily approximate the threshold value at ~30 days, 405 406 however this is an inflection point before CH₄ concentration begins to increase again in the

407 model to over 2 ppmy CH₄, a result not observed in the experimental data (brown curve in Fig. 4). Simple addition of CH₄ into the system at a rate similar to that observed in the anaerobic PF 408 microcosms (equivalent to $\sim 300 \text{ pmol CH}_4 \text{ d}^{-1}$) was insufficient to reproduce the observed 409 threshold alone and when combined with the ΔG_p of 287 kJ (mole of ATP)⁻¹ produced only a 410 411 marginal increase (orange curve in Fig. 4). Arbitrarily increasing the CH₄ input rates to 100 times the observed value in the anaerobic PF microcosms only created very large initial bursts of 412 CH₄ in the headspace that were not observed in the data and still failed to reproduce the 413 threshold observations (green curve in Fig. 4). 414

Aerobic CH₄ oxidation rates (Fig. S2a-c) ranged from 0.78 fmol CH₄ (kg-cryosol)⁻¹ s⁻¹ 415 (4°C 66% saturation) to 9.3 fmol kg-cryosol⁻¹ s⁻¹ (10°C 33% saturation). Reaction rates at 10°C 416 were 50 to 100% higher than those at 4°C. For the aerobic samples with 2 ppmv CH₄ the 417 calculated activation energies for the microbial CH₄ oxidation rates ranged from 101±57 kJ mol-418 CH_4^{-1} (pre-exponential factor = 1.7×10^{23} nmols of CH_4 g⁻¹ d⁻¹) at 33% saturation to 163 ± 159 kJ 419 mol-CH₄⁻¹ (pre-exponential factor = 2.6×10^{29} nmols of CH₄ g⁻¹ d⁻¹) at 100% saturation, with no 420 statistical difference between the treatments. The CH₄ concentrations in the microcosm water in 421 equilibrium with the 2 ppmv CH₄ headspace at 1.5 atmospheres were 7 and 6 nM at 4° and 10°C, 422 respectively. The modeled aerobic CH₄ oxidation rates decreased as the CH₄ concentrations 423 decreased (Fig. S2d-f). As it turns out these rates closely mimic the in situ rates since the original 424 0-10 cm samples were 100% saturated, the 30-40 cm samples were 33% saturated and the 60-70 425 and permafrost samples were 66% saturated with water. 426

427 Sequence abundance information from metagenomic data gathered from the 5 cm cryosol samples from the same site (Chauhan et al., 2014; Stackhouse et al., 2015) was used to estimate 428 that the total methanotrophic population should account for up to ~1% of the cellular community 429 with a higher proportion of Type II methanotrophs at the surface (~0.8% of total sequences) than 430 in the permafrost (~0.1% of total sequences). Total counts of bacterial, archaeal, and eukaryotic 431 cells determined by FISH averaged $9.8\pm5.2\times10^8$ cells g⁻¹, with bacteria accounting for $82\pm11\%$ 432 of cell counts and no statistical difference between depths (p = 0.46, Fig. S3). Assuming a cell 433 concentration of 9.8×10^8 cells g⁻¹ (Fig. S3) and a 0.8% abundance of Type II methanotrophs, this 434 would result in 7.8×10^6 Type II methanotrophic-cells g⁻¹. Assuming a cellular mass of 1×10^{-13} g 435 cell⁻¹ (Phelps et al., 1994) this would equate to ~ 0.8 mg-biomass kg-cryosol⁻¹. In order to capture 436

437 the lag phase of several days in the CH_4 oxidation rate observed in the 4°C treatments, the initial

438 metabolically active biomass had to be set to a lower value of 0.3 ng-biomass kg-cryosol⁻¹. At

439 the end of the simulation the Type II methanotroph biomass was 23 ng-biomass kg-cryosol⁻¹.

440 Assuming 7.8×10^6 atmospheric CH₄ oxidizing cells g⁻¹, the specific rate of CH₄ oxidation after

the lag phase was over ranged from 0.007 fmol CH₄ cell⁻¹ day⁻¹ (4°C 66% saturation) to 0.08

fmol CH₄ cell⁻¹ day⁻¹ (10°C 100% saturation) at 2 ppmv CH₄ (Fig. 5, Table 4).

443

444 Discussion

The higher CH₄ oxidation rates in the 0-10 cm cryosol samples were likely due to a larger 445 methanotrophic community (1.2% of total sequences) than at depth (<0.5% of total sequences), 446 modulated primarily by increases in Type II methanotrophs (Stackhouse et al., 2015). The lower 447 CH₄ oxidation rates with increasing water saturation for the 10°C incubations are likely due to 448 limitation of CH₄ and O₂ diffusion through water to the cells (Whalen & Reeburgh, 1996). The 449 lack of a correlation between the CH₄ oxidation rates with increasing water saturation at the 4°C 450 incubations may reflect the much lower rates of CH₄ oxidation and slightly higher O₂ 451 concentrations due to increased gas solubility, such that O_2 is no longer a limiting factor. 452

The Q₁₀ values for CH₄ oxidation in the 0-10 cm microcosms increased significantly with 453 decreasing water saturation (from 2.4±1.6 to 24.2±12.3), with these high values likely due to the 454 interaction of gas diffusion and temperature rather than simple temperature dependence of CH_4 455 oxidation alone (Table 3, Davidson & Janssens, 2006). Recent research indicates that the 456 457 methanotrophic response to warming conditions may be more acute at the lower temperature ranges of polar environments than the temperatures used in our incubation studies (Lupascu et 458 al., 2012; Lau et al., 2015). He et al. (2012) similarly observed increasing CH₄ oxidation rates 459 with increasing temperature in Arctic lake sediments (5-10 μ mol CH₄ g⁻¹ d⁻¹ at 4°C to 35-50 460 μ mol CH₄ g⁻¹ d⁻¹ at 21°C), as well as a shift in the methanotrophic community towards one with 461 an increased proportion of a Type II methanotroph (genus Methylocystis) in the surface as the 462 temperature increased. He et al. (2012) reported the cooccuring Type I methanotrophs shifted 463 464 from Methylomonas at 21°C to Methylobacter at 4°C along with the reduction of Methylocystis relative abundance to near zero. Rice paddies also show a strong preference for Type II 465

466 methanotrophs closer to the surface, with an increase by a factor of ~ 2 , positively correlating with pore water CH_4 concentrations (Macalady et al., 2002). Knoblauch et al. (2008) found that 467 Siberian polygon (0-5 cm) cryosols with an in situ MAT of -14°C have a CH₄ oxidation rate 468 maximum between 20 and 28°C (~55 \pm 20 nmol CH₄ gdw⁻¹ h⁻¹ compared to ~5 \pm 1 nmol CH₄ gdw⁻¹ 469 h^{-1} at 0°C). Interestingly, research from the same area in the Lena Delta, Siberia shows that CH₄ 470 oxidation rates exhibit maxima at the surface at high temperature and near the permafrost table at 471 low temperature, indicating that methanotrophic populations had adapted to their long-term in 472 situ temperatures (Liebner & Wagner, 2007). This trend was not observed in the current study. 473

474

The 50% CH₄ incorporation value found (Fig. 3) under all temperature and saturation 475 conditions in this study is greater than previously published experiments on methanotrophic 476 477 isolates reporting growth yields of 8 to 12 grams of biomass per mole of CH₄ oxidized for experiments where the CH₄ concentration ranged from 11,000 to 370,000 ppmv CH₄ and the 478 incubation temperatures ranged from 30°C to 45°C (Vary & Johnson, 1967; Leak & Dalton, 479 1986a). Our carbon use efficiency, however, is close to the 31 to 43% efficiency reported by 480 Roslev et al. (1997) using ¹⁴CH₄ incubation experiments on atmospheric CH₄ oxidizing soils at a 481 mixing ratios of 5 to 10 ppmv CH₄ and temperatures of 5 and 10°C. It is also comparable to the 482 31-49% carbon usage efficiency of limiting CH₄ by a Type I methanotrophy in culture (Leak & 483 Dalton, 1986a), and the 45-47% estimated carbon usage efficiency of a methanotroph using 484 nitrate as an N source (Leak & Dalton, 1986b). 485

Few studies have been conducted testing oxidation rates under atmospheric conditions of 486 1.8 ppmv CH₄. Experiments at CH₄ concentrations far in excess of those observed in nature may 487 influence observed rates by effectively measuring maximum oxidation rather than in situ 488 oxidation and may be selecting for low-affinity methanotrophs that are not performing the 489 majority of atmospheric CH₄ oxidation in the environment (Dunfield et al., 1999). At 490 atmospheric CH₄ concentrations of 1.8 ppmv and 4° to 10°C, AHI cellular rates (0.007 to 0.08 491 fmol of CH_4 cell⁻¹ day⁻¹) are up to 2-fold higher than those reported for 25°C incubations of 492 Methylocystis, Methylosinus, Methylocaldum, and Methylobacter at 1.8 ppmv CH₄ (Knief & 493 494 Dunfield, 2005; Baani & Liesack, 2008) (Fig. 2a, Table 5). These isolates had normalized

cellular CH₄ oxidation rates (cell⁻¹ day⁻¹) comparable to those of the same cultures grown at CH₄ 495 concentrations up to 1,000 ppmy. Interestingly, the studies with rates comparable to those from 496 497 this study are both derived from methanotroph batch cultures grown in media continuously shaken by means of a shake plate (Knief & Dunfield, 2005; Baani & Liesack, 2008). A higher 498 oxidation rate for shaken media would be expected since diffusion of CH₄ and O₂ would occur 499 much faster than through static soils (such as the incubation conditions presented here. The 500 501 cellular CH₄ oxidation rates reported by Liebner and Wagner (2007) for static microcosms using soil subsample (similar to this study) were more than an order of magnitude lower than the 502 results from this study, further highlighting the high affinity of AHI cryosol methanotrophs for 503 504 CH₄.

505 The factors determining the threshold concentration for CH₄ oxidation remain unclear. Both the low concentration of H_2 and the lack of increased methanogenesis from acetate 506 507 amendments argue against substantial hydrogenotrophic or acetoclastic methanogenesis occurring, implying that the steady CH_4 value is not a result of a balance between methanotrophy 508 509 and methanogenesis. Threshold values were lower at higher temperatures (0.1 to 0.5 ppmv CH₄ at 10°C vs 0.5 to 1.4 ppmv CH₄ at 4°C), and were not clearly correlated with water saturation 510 (Fig. S1). Field observations in Alaska have shown that at lower temperatures (2.5 to 5.5°C) 511 ambient CH_4 concentrations can be maintained at a threshold concentration of 0.1 ppmv 512 (Reeburgh et al., 1997), and that drawdown from up to 50 ppmv CH_4 to near-zero concentrations 513 (T=7°C) can occur on the timescale of tens of hours (Whalen & Reeburgh, 1990). Threshold 514 concentrations closer to those observed in this study were reported by Benstead et al. (1998) (0.5 515 ppmv CH₄ for Methylobacter and Methylosinus) and Jensen et al. (1998) (0.1 to 0.3 ppmv CH₄), 516 however these microcosms were run with additions of methanol and methanol and formate, 517 respectively. It is possible the methanotrophic strains present at AHI may be facultative 518 methanotrophs capable oxidizing atmospheric CH₄ while using simple carbon compounds other 519 520 than acetate as cosubstrates. The genome of acidophilic methanotroph Methylocella silvestris 521 $(0.53\pm0.24\%$ sequence abundance at 5 cm) shows it to be capable of using C₁, C₂ (particularly acetate), and C₃ compounds (Chen et al., 2010), whereas the genome of Methylococcus 522 523 capsulatus (0.10±0.01% sequence abundance at 5 cm) implies some metabolic flexibility and the 524 ability to use simple sugars for metabolic purposes (Ward et al., 2004). The addition of methanol 525 directly to forest soils in situ was found to decrease CH₄ oxidation potential as methanotrophs

using C_1 -compounds used that substrate instead (Jensen et al., 1998), a finding that may explain

527 why the acetate addition experiments in this study decreased CH₄ oxidation rates. Regardless, the

- variable threshold value may be a result of the methanotrophic community adjusting their kinetic
- properties in response to changing growth conditions and substrate availability (Koch, 1997;
- 530 Benstead et al., 1998).

Across all saturations, depths, and temperatures, AHI cryosols were able to oxidize CH₄ 531 532 at atmospheric concentrations, indicating the potential for these cryosols to act as sinks of atmospheric CH₄ with warming. Extrapolating the rates observed here to a surface atmospheric 533 CH₄ uptake yields fluxes ranging from 11.2 \pm 6.2 µmol CH₄ m⁻² d⁻¹ (4°C under fully saturated 534 conditions) to 72.6 \pm 14.4 µmol CH₄ m⁻² d⁻¹ (10°C under 33% water saturation conditions) 535 (compiled values shown in Table 5). Although these extrapolated fluxes assume that subsurface 536 pore gas contains atmospheric concentrations of CH₄, these values overlap the uptake rates 537 538 measured in the field with flux chambers (Stackhouse et al. 2015). The CH₄ oxidizing behavior of these cryosols, even under saturated conditions, is in stark contrast to cryosols from other 539 540 Arctic regions, which are generally reported to be CH₄ sources during full active layer thaw ranging from 40 nmol CH₄ m⁻² d⁻¹ to greater than 25 mmol CH₄ m⁻² d⁻¹ of emissions, across 541 temperatures ranging from 0.5°C to 24°C (Verville et al., 1998; Christensen et al., 2000; Wagner 542 et al., 2005; Mastepanov et al., 2008; Brummell et al., 2012; Lupascu et al., 2012). The 6 orders 543 of magnitude differences could be caused by variations in SOC availability, water saturation, 544 545 temperature and, also, the balance between methanogenic and methanotrophic activity.

CO₂ production was found to inversely correlate primarily with depth and was highest in 546 the 0-10 cm portion of the core where SOC content and biomass were the highest (Table 2). 547 Although the 66% and 100% saturation treatments were not statistically higher than the 33% 548 549 treatments, the range and upper bound of CO₂ emissions were larger, indicating that access to 550 carbon substrate may be limited by liquid phase transport in drier soils. Allison and Treseder (2008) observed drying of soils through warming to suppress CO₂ emissions from Alaskan soils 551 and to cause large shifts in the microbial and fungal SOC degrader communities. For the deeper 552 553 samples the CO₂ production rate appears relatively insensitive to both temperature and water saturation state, perhaps due to the low carbon quality of the $\sim 1\%$ SOC present at these depths 554 (Nadelhoffer et al., 1991; Jagadamma et al., 2014). 555

556 The per gram CO_2 production values from these experiments were used to extrapolate a CO_2 flux for these cryosols for the incubation temperatures and saturation conditions on a m⁻² 557 basis, assuming a soil density of 1.8 g cm^{-3} (Stackhouse et al., 2015) and assuming a deepened 558 active layer depth of 80 cm. The CO₂ production rates measured for the microcosms would 559 correspond to CO_2 emission flux of 52±5 mmol CO_2 m⁻² d⁻¹ under 100 % saturated conditions at 560 4°C and 41±7 mmol CO₂ m⁻² d⁻¹ at 33% saturation at 10°C (Table 5). Surface fluxes measured at 561 the same locations from which the cryosols samples were collected were $20\pm5 \text{ mmol CO}_2 \text{ m}^{-2} \text{ d}^{-1}$ 562 (soil T, 0.7 to 3.6°C) and 39 \pm 26 mmol CO₂ m⁻² d⁻¹ (soil T, 7.9 to 18.4°C), respectively (Allan et 563 al., 2014). This would be a moderate CO_2 source when compared to summertime active layer 564 fluxes from other sites in the Canadian and Alaskan Arctic, which range from -300 to 500 mmol 565 $CO_2 \text{ m}^{-2} \text{ d}^{-1}$ (Verville et al., 1998; Williams et al., 2000; Worthy & Levin, 2000; Brummell et al., 566 567 2012; Stackhouse et al., 2015).

568 Critical questions remain, however, as to the long-term duration of these responses to 569 temperature and saturation. Future studies examining the microbial response and geochemical 570 evolution of long-term changes in mineral cryosol permafrost systems will provide essential 571 information for modeling the Arctic carbon cycle through the rest of the century.

- 572
 - 72

573 Conclusion

The atmospheric CH₄ oxidation rates were measured in microcosms incubated at 4°C and 10°C along a 1 m depth profile and over a range of water saturation conditions for mineral cryosols containing Type I and Type II methanotrophs from Axel Heiberg Island (AHI), Nunavut, Canada. Our research indicates that for cryosols derived from Axel Heiberg Island the following parameters impact CH₄ uptake and CO₂ emission rates:

Increasing water saturation, while limiting CH₄ oxidation in cryosols, does not
 eliminate it or result in net CH₄ emissions. Increasing water saturation was not found
 to have a significant effect on CO₂ emissions from cryosols.

582 583

584

CH₄ oxidation was increased by roughly a factor of 2 for samples from the active layer (0-70 cm) by a shift in temperature from 4°C to 10°C. A similar trend held true for CO₂ emissions only for the top 10 cm, where SOC content was the highest.

- 5853. CH4 oxidation from Axel Heiberg cryosols is able to occur at a higher cellular rate586than many other Arctic sites when accounting for temperature and CH4 concentration587of incubation, indicating that methanotrophs at this site have a high affinity for CH4.
- 588Biomass incorporation of oxidized CH4 was shown at 10 ppmv CH4, suggesting that589AHI methanotrophs may be able to grow on CH4 at this concentration.

The cryosols exhibited net consumption of 1.8 ppmv CH₄ under all conditions, including 590 during anaerobic incubations. Methane oxidation rates increased with temperature and decreased 591 with increasing water saturation at 10°C and depth, exhibiting the highest rates at 10°C and 33% 592 saturation at 5 cm depth (260 \pm 60 pmol CH₄ gdw⁻¹ d⁻¹). The CH₄ oxidation rates for the 4°C 593 incubations, however, did not vary significantly with water saturation. Extrapolation of the CH₄ 594 oxidation rates to the field yields net CH₄ uptake rates ranging from 11 to 73 μ mol CH₄ m⁻² d⁻¹, 595 which are comparable to field measurements. Stable isotope mass balance indicates ~50% of the 596 oxidized CH_4 is incorporated into the methanotrophic biomass at a CH_4 concentration of only 10 597 598 ppmv, regardless of temperature or saturation. Future atmospheric CH₄ uptake rates at AHI with increasing temperatures will be determined by the interplay of increasing CH₄ oxidation rates 599 versus water saturation and the depth to the water table during summer thaw. CO_2 production 600 was found to inversely correlate primarily with depth and was highest in the 0-10 cm portion of 601 the core where SOC content and biomass were the highest. Thawing of high SOC permafrost 602 cryosols can lead to rapid changes in the microbial community and metabolic pathways 603 (Mackelprang et al., 2011) and in the degradation of SOC (Coolen et al., 2011), affecting the 604 substrates available for metabolism. Although the availability of soil C and N stores has been 605 shown to enhance CH₄ production in high SOC Alaskan soils (Waldrop et al., 2010) the direct 606 addition of acetate had no effect on either net CH₄ oxidation or production in this study (Fig. 2, 607 Table 3). 608

609

610 Acknowledgements

- 611 This research was funded by U. S. Department of Energy Office of Science, Office of Biological
- and Environmental Research, Genomic Science Program (No. DE-SC0004902) to T.C.O. and
- 613 logistical support was provided by the Canadian Polar Continental Shelf Program to L. Whyte of
- 614 McGill University. Field assistance during sample collection was provided by G. Lamarche-
- 615 Gagnon, R. Wilhelm, and N. Mykytczuk of McGill University.
- 616 References
- Adachi K, Ohtsuka T, Nakatsubo T, Koizumi H (2006) The methane flux along topographical
- gradients on a glacier foreland in the High Arctic, Ny-Ålesund, Svalbard. Polar Biosci 20, 131–
 139.
- Allan J, Ronholm J, Mykytczuk NS, Greer CW, Onstott TC, Whyte LG (2014) Methanogen
- 621 community composition and rates of methane consumption in Canadian High Arctic permafrost
- 622 soils. Environmental Microbiology Reports.
- Allison SD, Treseder KK (2008) Warming and drying suppress microbial activity and carbon
- 624 cycling in boreal forest soils. Global Change Biology **14**, 2898–2909.
- Allison SD, Wallenstein MD, Bradford MA (2010) Soil-carbon response to warming dependent
 on microbial physiology. Nature Geoscience 3, 336–340.
- 627 Ambus P, Andersen BL, Kemner M, Sørensen B, Wille J (2002) Natural carbon isotopes used to
- study methane consumption and production in soil. Isotopes in environmental and health studies
 38, 149–157.
- Baani M, Liesack W (2008) Two isozymes of particulate methane monooxygenase with different
- 631 methane oxidation kinetics are found in Methylocystis sp. strain SC2. Proceedings of the
- 632 National Academy of Sciences 105, 10203–8.
- 633 Bäckstrand K, Crill PM, Jackowicz-Korczyński M, Mastepanov M, Christensen TR, Bastviken D
- 634 (2010) Annual carbon gas budget for a subarctic peatland, Northern Sweden. Biogeosciences 7,
- 635 95–108.
- Bárcena TG, Finster KW, Yde JC (2011) Spatial Patterns of Soil Development, Methane
- 637 Oxidation, and Methanotrophic Diversity along a Receding Glacier Forefield, Southeast
- Greenland. Arctic, Antarctic, and Alpine Research 43, 178–188.

- Bender M, Conrad R (1992) Kinetics of CH 4 oxidation in oxic soils exposed to ambient air or
- high CH 4 mixing ratios. FEMS Microbiology Letters **101**, 261–269.
- 641 Benstead J, King GM, Williams HG (1998) Methanol promotes atmospheric methane oxidation
- by mathanotrophic cultures and soils **64**, 1091–1098.
- Bethke CM (2008) Geochemical and Biogeochemical Reaction Modeling, 2nd editio. University
 of Illinois, Urbana, Illinois.
- Bratbak G, Dundas I (1984) Bacterial dry matter content and biomass estimations. Applied and
 environmental microbiology 48, 755–757.
- 647 Brummell ME, Farrell RE, Hardy SP, Siciliano SD (2014) Greenhouse gas production and
- 648 consumption in High Arctic deserts. Soil Biology and Biochemistry 68, 158–165.
- 649 Brummell ME, Farrell RE, Siciliano SD (2012) Greenhouse gas soil production and surface
- 650 fluxes at a high arctic polar oasis. Soil Biology and Biochemistry **52**, 1–12.
- 651 Chauhan A, Layton AC, Vishnivetskaya TA, Williams D, Pfiffner SM, Rekepalli B (2014)
- 652 Metagenomes from Thawing Low-Soil-Organic-Carbon Mineral Cryosols and Permafrost of the
- 653 Canadian High Arctic. Genome Announcements 2, 1–2.
- 654 Chen Y, Crombie A, Rahman MT, Dedysh SN, Liesack W, Stott MB, Alam M, Theisen AR,
- 655 Murrell JC, Dunfield PF (2010) Complete genome sequence of the aerobic facultative
- 656 methanotroph Methylocella silvestris BL2. Journal of bacteriology **192**, 3840–1.
- 657 Christensen T, Friborg T, Sommerkorn M, Kaplan J, Illeris L, Soegaard H, Nordstroem C,
- Jonasson S (2000) Trace gas exchange in a high Arctic valley: 1. Variations in CO2 and CH4
- Flux between tundra vegetation types. Global Biogeochemical Cycles 14, 701–713.
- 660 Christensen T, Jonasson S, Callaghan T, Havstrom M (1995) Spatial variation in high-latitude
- 661 methane flux along a transect across Siberian and European tundra environments. Journal of
- 662 Geophysical Research **100**, 21035–21045.
- 663 Coolen MJL, Giessen J van de, Zhu EY, Wuchter C (2011) Bioavailability of soil organic matter
- and microbial community dynamics upon permafrost thaw. Environmental microbiology **13**,
- 665 2299–314.

- 666 Davidson E, Janssens I (2006) Temperature sensitivity of soil carbon decomposition and
- 667 feedbacks to climate change. Nature **440**, 165–73.
- 668 Degelmann DM, Borken W, Drake HL, Kolb S (2010) Different atmospheric methane-oxidizing
- 669 communities in European beech and Norway spruce soils. Applied and environmental
- 670 microbiology **76**, 3228–35.
- 671 Dlugokencky EJ, Lang PM, Crotwell AM, Masarie KA, Crotwell M. (2014) Atmospheric
- 672 Methane Dry Air Mole Fractions from the NOAA ESRL Carbon Cycle Cooperative Global Air
- 673 Sampling Network, 1983-2013 2014–06–24.
- 674 Dunfield P, Liesack W, Henckel T, Knowles R, Conrad R (1999) High-affinity methane
- 675 oxidation by a soil enrichment culture containing a type II methanotroph. Applied and
- environmental microbiology **65**, 1009–1014.
- 677 Emmerton CA, Louis VL St., Lehnherr I, Humphreys ER, Rydz E, Kosolofski HR (2014) The
- 678 net exchange of methane with high Arctic landscapes during the summer growing season.
- Biogeosciences Discussions 11, 1673–1706.
- 680 Girguis P, Orphan V, Hallam S, DeLong E (2003) Growth and methane oxidation rates of
- anaerobic methanotrophic archaea in a continuous-flow bioreactor. Applied and environmental
 microbiology 69, 5472–5482.
- 683 Graef C, Hestnes AG, Svenning MM, Frenzel P (2011) The active methanotrophic community in
- a wetland from the High Arctic. Environmental microbiology reports 3, 466-72.
- 685 Graham DE, Wallenstein MD, Vishnivetskaya T a, Waldrop MP, Phelps TJ, Pfiffner SM,
- 686 Onstott TC, Whyte LG, Rivkina EM, Gilichinsky DA, Elias DA, Mackelprang R, VerBerkmoes
- 687 NC, Hettich RL, Wagner D, Wullschleger SD, Jansson JK (2012) Microbes in thawing
- 688 permafrost: the unknown variable in the climate change equation. The ISME journal **6**, 709–12.
- 689 Grosse G, Harden J, Turetsky M, McGuire AD, Camill P, Tarnocai C, Frolking S, Schuur E a G,
- Jorgenson T, Marchenko S, Romanovsky V, Wickland KP, French N, Waldrop M, Bourgeau-
- 691 Chavez L, Striegl RG (2011) Vulnerability of high-latitude soil organic carbon in North America
- to disturbance. Journal of Geophysical Research: Biogeosciences **116**, 1–23.
- Hayes DJ, Kicklighter DW, McGuire AD, Chen M, Zhuang Q, Yuan F, Melillo JM,

- Wullschleger SD (2014) The impacts of recent permafrost thaw on land–atmosphere greenhouse
- 695 gas exchange. Environmental Research Letters 9, 045005.
- He R, Wooller MJ, Pohlman JW, Quensen J, Tiedje JM, Leigh MB (2012) Shifts in identity and
- 697 activity of methanotrophs in arctic lake sediments in response to temperature changes. Applied
- and environmental microbiology **78**, 4715–23.
- Hill GB, Henry GHR (2011) Responses of High Arctic wet sedge tundra to climate warming
 since 1980. Global Change Biology 17, 276–287.
- 701 Hugelius G, Strauss J, Zubrzycki S, Harden JW, Schuur EAG, Ping CL, Schirrmeister L, Grosse
- G, Michaelson GJ, Koven CD, O'Donnell JA, Elberling B, Mishra U, Camill P, Yu Z, Palmtag J,
- Kuhry P (2014) Improved estimates show large circumpolar stocks of permafrost carbon while
- 704 quantifying substantial uncertainty ranges and identifying remaining data gaps. Biogeosciences
- 705 Discussions **11**, 4771–4822.
- Jagadamma S, Mayes M a., Steinweg JM, Schaeffer SM (2014) Substrate quality alters the
- microbial mineralization of added substrate and soil organic carbon. Biogeosciences 11, 4665–
 4678.
- Jensen S, Priemé A, Bakken L (1998) Methanol improves methane uptake in starved
- methanotrophic microorganisms. Applied and Environmental Microbiology **64**, 1143–1146.
- 711 Jetten MSM, Stams AJM, Zehnder AJB (1990) Acetate threshold values and acetate activating
- enzymes in methanogenic bacteria. FEMS Microbiology Letters **73**, 339–344.
- Jin Q, Bethke CM (2003) A new rate law describing microbial respiration. Applied and
- 714 Environmental Microbiology **69**, 2340–2348.
- Joergensen L (1985) The methane mono-oxygenase reaction system studied in vivo by
- membrane-inlet mass spectrometry. The Biochemical Journal **225**, 441–8.
- Jorgenson MT, Shur YL, Pullman ER (2006) Abrupt increase in permafrost degradation in
- 718 Arctic Alaska. Geophysical Research Letters **33**, L02503.
- 719 Knief C, Dunfield PF (2005) Response and adaptation of different methanotrophic bacteria to
- low methane mixing ratios. Environmental microbiology **7**, 1307–17.

- 721 Knoblauch C, Zimmermann U, Blumenberg M, Michaelis W, Pfeiffer E (2008) Methane
- turnover and temperature response of methane-oxidizing bacteria in permafrost-affected soils of
- northeast Siberia. Soil Biology and Biochemistry **40**, 3004–3013.
- Koch A (1997) Microbial physiology and ecology of slow growth. Microbiology and molecular
- 725 biology reviews **61**, 305–318.
- Kolb S (2009) The quest for atmospheric methane oxidizers in forest soils. Environmental
- microbiology reports 1, 336–46.
- 728 Kolb S, Knief C, Dunfield PF, Conrad R (2005) Abundance and activity of uncultured
- methanotrophic bacteria involved in the consumption of atmospheric methane in two forest soils.
- 730 Environmental Microbiology 7, 1150–1161.
- 731 Koven CD, Ringeval B, Friedlingstein P, Ciais P, Cadule P, Khvorostyanov D, Krinner G,
- 732 Tarnocai C (2011) Permafrost carbon-climate feedbacks accelerate global warming. Proceedings
- of the National Academy of Sciences of the United States of America **108**, 14769–74.
- Lau MCY, Stackhouse BT, Layton AC, Chauhan A, Vishnivetskaya TA, Chourey K, Ronholm J,
- 735 Mykytczuk NCS, Bennett PC, Lamarche-Gagnon G, Burton N, Pollard WH, Omelon CR,
- 736 Medvigy DM, Hettich RL, Pfiffner SM, Whyte LG, Onstott TC (2015) An active atmospheric
- 737 methane sink in high Arctic mineral cryosols. The ISME Journal.
- T38 Lawrence DM, Slater AG, Romanovsky VE, Nicolsky DJ (2008) Sensitivity of a model
- 739 projection of near-surface permafrost degradation to soil column depth and representation of soil
- organic matter. Journal of Geophysical Research **113**, F02011.
- Leak D, Dalton H (1986a) Growth yields of methanotrophs 1. Applied microbiology and
 biotechnology 23, 470–476.
- Leak D, Dalton H (1986b) Growth yields of methanotrophs 2: A theoretical analysis. Applied
 microbiology and biotechnology 23, 477–481.
- Liebner S, Wagner D (2007) Abundance, distribution and potential activity of methane oxidizing
- bacteria in permafrost soils from the Lena Delta, Siberia. Environmental microbiology 9, 107–
 17.
- Liebner S, Zeyer J, Wagner D, Schubert C, Pfeiffer E-M, Knoblauch C (2011) Methane

- oxidation associated with submerged brown mosses reduces methane emissions from Siberian
- polygonal tundra. Journal of Ecology **99**, 914–922.
- 751 Lupascu M, Wadham J, Hornibrook E, Pancost R (2012) Temperature Sensitivity of Methane
- 752 Production in the Permafrost Active Layer at Stordalen, Sweden: A Comparison with Non-
- permafrost Northern Wetlands. Arctic, Antarctic, and Alpine Research 44, 469–482.
- 754 Macalady JL, McMillan AMS, Dickens AF, Tyler SC, Scow KM (2002) Population dynamics of
- type I and II methanotrophic bacteria in rice soils. Environmental Microbiology **4**, 148–157.
- 756 Mackelprang R, Waldrop MP, DeAngelis KM, David MM, Chavarria KL, Blazewicz SJ, Rubin
- EM, Jansson JK (2011) Metagenomic analysis of a permafrost microbial community reveals a
- rapid response to thaw. Nature **480**, 368–71.
- 759 Martineau C, Pan Y, Bodrossy L, Yergeau E, Whyte LG, Greer CW (2014) Atmospheric
- methane oxidizers are present and active in Canadian high Arctic soils. FEMS microbiology
- 761 ecology 1–13.
- 762 Mastepanov M, Sigsgaard C, Dlugokencky EJ, Houweling S, Ström L, Tamstorf MP,
- 763 Christensen TR (2008) Large tundra methane burst during onset of freezing. Nature 456, 628–
 764 30.
- Montzka S, Dlugokencky E, Butler J (2011) Non-CO2 greenhouse gases and climate change.
 Nature 476, 43–50.
- Nadelhoffer K, Giblin A, Shaver G, Laundre J (1991) Effects of temperature and substrate
- quality on element mineralization in six arctic soils. Ecology **72**, 242–253.
- 769 Ohtsuka T, Adachi M, Uchida M, Nakatsubo T (2006) Relationships between vegetation types
- and soil properties along a topographical gradient on the northern coast of the Brøgger Peninsula,
- 771 Svalbard. Polar Bioscience **19**, 63–72.
- 772 Oremland R, Miller L, Whiticar M (1987) Sources and flux of natural gases from Mono Lake,
- 773 California. Geochimica et Cosmochimica Acta **51**, 2915–2929.
- 774 Osterkamp TE (2007) Characteristics of the recent warming of permafrost in Alaska. Journal of
- 775 Geophysical Research **112**, F02S02.

- Phelps TJ, Murphy EM, Pfiffner SM, White DC (1994) Comparison between geochemical and
- biological estimates of subsurface microbial activities. Microbial ecology **28**, 335–49.
- 778 Reeburgh WS, Hirsch a I, Sansone FJ, Popp BN, Rust TM (1997) Carbon kinetic isotope effect
- accompanying microbial oxidation of methane in boreal forest soils. Geochimica et
- 780 Cosmochimica Acta **61**, 4761–4767.
- 781 Roslev P, Iversen N, Henriksen K (1997) Oxidation and assimilation of atmospheric methane by
- soil methane oxidizers. Applied and environmental microbiology **63**, 874–80.
- 783 Rowlands DJ, Frame DJ, Ackerley D, Aina T, Booth BBB, Christensen C, Collins M, Faull N,
- Forest CE, Grandey BS, Gryspeerdt E, Highwood EJ, Ingram WJ, Knight S, Lopez A, Massey
- N, McNamara F, Meinshausen N, Piani C, Rosier SM, Sanderson BM, Smith LA, Stone DA,
- 786 Thurston M, Yamazaki K, Hiro Yamazaki Y, Allen MR (2012) Broad range of 2050 warming
- from an observationally constrained large climate model ensemble. Nature Geoscience 5, 256–
 260.
- Sachs T, Wille C, Boike J, Kutzbach L (2008) Environmental controls on ecosystem-scale CH 4
 emission from polygonal tundra in the Lena River Delta, Siberia. Journal of Geophysical
- 791 Research **113**, G00A03.
- 792 Schuur EAG, Bockheim J, Canadell JG, Euskirchen E, Field CB, Goryachkin S V., Hagemann S,
- 793 Kuhry P, Lafleur PM, Lee H, Mazhitova G, Nelson FE, Rinke A, Romanovsky VE,
- 794 Shiklomanov N, Tarnocai C, Venevsky S, Vogel JG, Zimov SA (2008) Vulnerability of
- Permafrost Carbon to Climate Change: Implications for the Global Carbon Cycle. BioScience 58,
 796 701.
- 797 Shindell DT, Faluvegi G, Koch DM, Schmidt GA, Unger N, Bauer SE (2009) Improved
- attribution of climate forcing to emissions. Science (New York, N.Y.) **326**, 716–718.
- 799 Stackhouse BT, Vishnivetskaya TA, Layton A, Chauhan A, Pfiffner S, Mykytczuk NC, Sanders
- 800 R, Whyte LG, Hedin L, Saad N, Myneni S, Onstott TC (2015) Simulated spring thaw of
- 801 permafrost from mineral cryosol reveals increase in CO2 emissions and an atmospheric CH4
- sink. Journal of Geophysical Research: Biogeosciences In press, n/a–n/a.
- 803 Temkin MI (1963) Kinetics of stationary reactions. Proceedings of the USSR Academy of

804 Sciences **152**, 156–159.

Thauer RK, Jungermann K, Decker K (1977) Energy conservation in chemotrophic anaerobic
bacteria. Bacteriological reviews 41, 100–180.

Vary P, Johnson M (1967) Cell yields of bacteria grown on methane. Applied microbiology 15,
1473–1478.

809 Verville J, Hobbie S, Chapin III F, Hooper D (1998) Response of tundra CH4 and CO2 flux

tomanipulation of temperature and vegetation. Biogeochemistry 4, 215–235.

811 Vishnivetskaya T a, Layton AC, Lau MCY, Chauhan A, Cheng KR, Meyers AJ, Murphy JR,

812 Rogers AW, Saarunya GS, Williams DE, Pfiffner SM, Biggerstaff JP, Stackhouse BT, Phelps TJ,

813 Whyte L, Sayler GS, Onstott TC (2014) Commercial DNA extraction kits impact observed

microbial community composition in permafrost samples. FEMS microbiology ecology 87, 217–
30.

816 Wagner D, Kobabe S, Pfeiffer E-M, Hubberten H-W (2003) Microbial controls on methane

fluxes from a polygonal tundra of the Lena Delta, Siberia. Permafrost and Periglacial Processes
14, 173–185.

819 Wagner D, Lipski A, Embacher A, Gattinger A (2005) Methane fluxes in permafrost habitats of

the Lena Delta: effects of microbial community structure and organic matter quality.

821 Environmental microbiology 7, 1582–92.

Waldrop MP, Wickland KP, White R, Berhe a. a., Harden JW, Romanovsky VE (2010)

823 Molecular investigations into a globally important carbon pool: Permafrost-protected carbon in

Alaskan soils. Global Change Biology 16, 2543–2554.

825 Walter KM, Zimov S, Chanton JP, Verbyla D, Chapin FS (2006) Methane bubbling from

Siberian thaw lakes as a positive feedback to climate warming. Nature 443, 71–5.

827 Ward N, Larsen Ø, Sakwa J, Bruseth L, Khouri H, Durkin a S, Dimitrov G, Jiang L, Scanlan D,

828 Kang KH, Lewis M, Nelson KE, Methé B, Wu M, Heidelberg JF, Paulsen IT, Fouts D, Ravel J,

829 Tettelin H, Ren Q, Read T, DeBoy RT, Seshadri R, Salzberg SL, Jensen HB, Birkeland NK,

830 Nelson WC, Dodson RJ, Grindhaug SH, Holt I, Eidhammer I, Jonasen I, Vanaken S, Utterback

T, Feldblyum T V, Fraser CM, Lillehaug JR, Eisen J a (2004) Genomic insights into

- methanotrophy: the complete genome sequence of Methylococcus capsulatus (Bath). PLoS
- biology **2**, e303.
- 834 Westermann P, Ahring BK, Mah R a (1989) Threshold acetate concentrations for acetate
- catabolism by aceticlastic methanogenic bacteria. Applied and environmental microbiology 55,
- 836 514–515.
- 837 Whalen S, Reeburgh W (1992) Interannual variations in tundra methane emission: A 4-year time
- series at fixed sites. Global Biogeochemical Cycles 6, 139–159.
- 839 Whalen S, Reeburgh W (1996) Moisture and temperature sensitivity of CH4 oxidation in Boreal
- soils. Soil Biology and Biochemistry **28**, 1271–1281.
- 841 Whalen SC, Reeburgh WS (1990) Consumption of atmospheric methane by tundra soils. Nature



- 843 Williams M, Eugster W, Rastetter E, McFadden J, III FC (2000) The controls on net ecosystem
- productivity along an Arctic transect: a model comparison with flux measurements. Global
- 845 Change Biology **6**, 116–126.
- Worthy D, Levin I (2000) Evidence for a link between climate and northern wetland methane
 emissions. Journal of Geophysical Research 105, 4031–4038.
- 848
- 849
- 850 Fig. legends
- 851 Fig. 1.
- Fig. 1. The average of triplicate CH₄ oxidation rates at 4°C and 10°C as a function of water
 saturation for (a) 0-10 cm, (b) 30-40 cm, (c) 60-70 cm, (d) 70-80 cm. Error bars represent ± 1
 S.D.

856

857 Fig. 2

858

Fig. 2. CH_4 oxidation rates and CO_2 emission rates for microcosms incubated at 4.5°C and examining light versus dark response and response to acetate, bicarbonate and water amendments. Negative values indicate uptake CH_4 into the soil. Error bars represent ± 1 S.D.

862 Fig. 3

Fig. 3 Caption: (a) 4°C and (b) 10°C comparison between observed and expected CH₄ loss based on direct measurements and isotopic tracer methods. 4°C (m = 0.47 ± 0.01 , R² = 0.99); 10°C (m = 0.56 ± 0.04 , R² = 0.97). Error bars represent ± 1 S.D.

- 867
- 868 Fig. 4

Fig. 4. The results of different modeling components overlain on the experimental results from the 4°C 33% saturation microcosm. MT stands for the methanotrophy only model. MT + AcMG stands for the methanotrophy model with acetoclastic methanogenesis added. MT + 300 pmol $CH_4 d^{-1}$ is the methanotrophy model with 300 pmol CH_4 added into the system daily. MMT + 30 nmol $CH_4 d^{-1}$ represents a methanotrophy model with a lowered CH_4 oxidation activation energy and 30 nmol CH_4 added into the system daily. Error bars represent ± 1 S.D.

- 876
- 877 Fig. 5
- 878
- Fig. 5. Cellular oxidation rates normalized to concentration of CH₄ as a function of (a) CH₄
 concentration and (b) temperature compared to values calculated from the literature.

- 881 Methanotrophs being compared are Methylocystis sp. Strain SC2 (Type II) (Baani & Liesack,
- 2008), Methylocystis and Methylosinus (Type II) and Methylocaldum and Methlyobacter (Type
- 883 II) (Knief & Dunfield, 2005), Anaerobic CH₄ oxidation by marine sediments from CH₄ seep
- (ANME-2c) (Girguis et al., 2003), and an environmental mix of Type I and II from 30 cm depth
- polygon soil from Lena Delta, Russia (Liebner & Wagner, 2007).
- 886
- 887 Fig. S1

888

Fig. S1. Measured (diamonds) and modeled (line) CH_4 headspace depletion during the course of soil incubations. Error bars represent ± 1 S.D. Where no error bars are visible the symbol is

- 891 larger than the error.
- 892 893 Fig. S2
- 894

Fig. S2. Reaction rates and thermodynamic potential factor (F_T , eq. 8) of surface soil (0-10 cm) microcosm conditions and permafrost soil (70-80 cm) microcosm conditions.

897

- 898 Fig. S3 899
- 900 Fig. S3. Estimation of cellular abundances of a depth profile by the FISH.

901 Tables

Table 1. Soil organic carbon content and in situ CH₄ fluxes from cryosol soil surfaces from sites across the Arctic.

Location	SOC (%)	CH ₄ Oxidation Rate	Reference
Lena Delta, Siberia	2-18	2.2 - 7.5 mmol $CH_4 m^{-2} d^{-1}$	(Wagner et al., 2003, 2005; Sachs et al., 2008)
Samoylov, Siberia	20-39	-835 nmol $CH_4 \text{ m}^{-2} \text{ d}^{-1} *$	(Knoblauch et al., 2008)
Stordalen, Sweden	24-59	$0.04 - 9.9 \text{ mmol CH}_4 \text{ m}^{-2} \text{ d}^{-1}$	(Bäckstrand et al., 2010; Lupascu et al., 2012)
Zackenberg Valley, Greenland	NR	1.4 - 25.5 mmol $CH_4 m^{-2} d^{-1}$	(Mastepanov et al., 2008)
Ellesmere Island, Nunavut	15-22	-1.5 - 3.1 μ mol CH ₄ m ⁻² d ⁻¹	(Hill & Henry, 2011; Brummell et al., 2012)
Unalaska Island, Alaska	16-36	-0.17 mmol $CH_4 m^{-2} d^{-1}$	(Whalen & Reeburgh, 1990)
Siberian polygons and bogs	6-95	$0.14 - 2.9 \text{ mmol CH}_4 \text{ m}^{-2} \text{ d}^{-1}$	(Christensen et al., 1995)
Spitsbergen, Svalbard	3-26	-0.50.29 nmol $CH_4 g^{-1} d^{-1}$	(Ohtsuka et al., 2006; Graef et al., 2011)
Ammassalik Island, Greenland	<1	-2.1 nmol $CH_4 g^{-1} d^{-1}$	(Bárcena et al., 2011)
Samoylov Island, Siberia	NR	-0.1 - 1.4 mmol $CH_4 m^{-2} d^{-1}$	(Liebner et al., 2011)
Axel Heiberg Island, Nunavut	1-5	-9.4 μ mol CH ₄ m ⁻² d ⁻¹	(Allan et al., 2014)

903

* indicates that CH_4 fluxes were determined from experiments using higher than atmospheric CH_4 concentrations. Negative values indicate CH_4 uptake by the soils.

Aerobic	0-10 cm		30_40	cm	60-70	cm	70 80 cm		
1.8 ppm CH_4	0-10 cm		30-40 cm		00-70	, cm	70-80 cm		
	CO ₂ (nmo	$CO_2 (nmol gdw^{-1} d^{-1}) CO_2 (nmol gdw^{-1} d^{-1}) CO_2 (nmol gdw^{-1} d^{-1})$			$CO_2 \text{ (nmol gdw}^{-1} d^{-1}\text{)}$				
% Saturation	4° C	10° C	4° C	10° C	4° C	10° C	4° C	10° C	
33	106 ± 26	98 ± 17	22*	13 ± 3	20 ± 1	37 ± 5	19 ± 7	21 ± 5	
66	142 ± 16	222 ± 77	19 ± 1	22 ± 6	42 ± 5	37 ± 2	37 ± 2	38 ± 8	
100	125 ± 14	235 ± 150	16 ± 2	21 ± 2	44 ± 2	40 ± 5	34 ± 2	36 ± 18	
		_							
	0-10 cm		70-80 cm		13CH ₄ microcosms	0-1	0 cm		
		$CO_2 \text{ (nmol } g^{-1} d^{-1}\text{)}$		-		CO ₂ (nn	nol $g^{-1} d^{-1}$)	-	
Treatment	AL Light	AL Dark	PF Dark (anaerobic)		% Saturation	4° C	10° C		
in situ*	252 ± 110	105 ± 23	29 ± 6	-	33	63 ± 26	55 ± 3	-	
Water*	420 ± 63	179 ± 53	34 ± 28		66	75 ± 14	95 ± 15		
Acetate*	413 ± 84	111 ± 9	3 ± 2		100	71 ± 10	151 ± 96		
Bicarbonate*	105 ± 15	156 ± 44	7 ± 1						
Aerobic	0.1		20.40		<i>(</i> 0. 7 0		70.4		
1.8 ppm CH ₄	0-1	to em	30-40) cm	00-70) cm	/0-2	so cm	
	CH ₄ (pm	ol $\mathrm{gdw}^{-1} \mathrm{d}^{-1}$)	CH ₄ (pmol	gdw ⁻¹ d ⁻¹)	CH ₄ (pmol	gdw ⁻¹ d ⁻¹)	CH ₄ (pm	(pmol gdw ⁻¹ d ⁻¹)	
	(k (g	$g^{-1} d^{-1}))$	(k (g ⁻¹	d ⁻¹))	(k (g ⁻¹	d ⁻¹))	(k (g	⁻¹ d ⁻¹))	
% Saturation	4° C	10° C	4° C	10° C	4° C	10° C	4° C	10° C	
33	-44 ± 20	-260 ± 60	-2.9*	-42.4 ± 2.7	0.3 ± 1.9	-7.3 ± 1.8	-7.2 ± 2.7	-1.6 ± 9.0	
	(-0.011±0.006)	(-0.036±0.005)	(-0.0002)	(-0.0013±0.0004)	(-0.0001±0.0001)	(0.0001±0.0001)	(-0.0002±0.0001)	(-0.0005±0.0003)	
66	-16 ± 12	-53 ± 18	$\textbf{-3.2}\pm1.7$	-15.7 ± 2.6	$\textbf{-1.2}\pm1.9$	$\textbf{-8.0}\pm0.4$	$\textbf{-10.0} \pm 3.4$	0.1 ± 2.2	
	(-0.003±0.002)	(-0.004±0.002)	(-0.001±0.0007)	(-0.0012±0.0003)	(-0.00001±0.0001)	(0.0004 ± 0.0001)	(-0.0003±0.0003)	(-0.0001±0.0001)	
100	-34 ± 19	-54 ± 18	-7.5 ± 4.5	-18.7 ± 11.3	$\textbf{-0.7} \pm 1.0$	-6.0 ± 2.0	$\textbf{-6.0} \pm 0.4$	$\textbf{-6.6} \pm 10.8$	
	(-0.002±0.001)	(-0.005±0.004)	(-0.0015±0.0009)	(-0.0017±0.0001)	(0.0001±0.0001)	(0.0002±0.0001)	(-0.0001±0.0002)	(-0.001±0.0004)	
	0.1		70.90			0.10	***		
			70-80 cm	-			1 -1 1-1**	-	
Turneturent	AT TILL	CH ₄ (pmol g [•] d [•])**			Cotore time	CH ₄ (pmo	ol g · d ·)**		
Ireatment	AL Light	AL Dark	PF Dark (anaerobic)	_	Saturation	4° C	10° C	-	
in situ*	-288 ± 123	-411 ± 70	-6 ± 1	D	33	-120 ± 130	-710 ± 300		
W/	(-0.015±0.004)	(-0.022 ± 0.002)	NA 1 + 2	Direct	66	-10 ± 10	-350 ± 50		
water*	-105 ± 11	-119 ± 2	-4 ± 2		100	-130 ± 90	-450 ± 24		
A catata*	(-0.0035 ± 0.0003)	(-0.0002 ± 0.0001)	NA 2 + 3	Isotope	55	-60 ± 60	-420 ± 20		
Actuac	-30 ± 2	(0.0082 ± 0.0005)	-2 ± 5	Isotope	100	-10 ± 10 70 ± 40	-240 ± 50		
Bicarbonate*	-112 + 1	-129 + 4	0 + 1		100	-70 ± 40	-290 ± 150		
Dicarbonate	(-0.0058+0.0001)	(-0.0067+0.0001)	NA						
	(
Anaerobic					Anaer	obic			
2.6 ppm CH₄	70-	80 cm			10 ppm	n CH4	70-8	30 cm	
	CH ₄ (pmo)	l gfw ⁻¹ d ⁻¹)**			· rr	-	CH4 (pmc	d^{-1} d ⁻¹)	
	4° C	10° C					4° C	10° C	
	-8.6±0.9	-8.6±1.4					BDL	BDL	
	(-0.0005±0.0001)	(-0.0006±0.00001)							
	,	(

Table 2. CO₂ emission and CH₄ oxidation rates from aerobic and anaerobic microcosms.

907

908 *Rates represent averages of 3 microcosms, except those marked by a * . Error shown is ± 1 S.D.

912

- 909 ** Negative values represent oxidation of CH₄.
- 910 ***For the isotopic versus direct CH_4 oxidation measurements the 10°C CH_4 rates are calculated
- 911 from day 18 values. The $4^{\circ}C$ CH₄ rates are calculated from 38 day values.

anusc J V Aut

914 CH_4 emissions.

		_											
			0-10 cm			30-40 cm			60-70 cm			70-80 cm	
		Water			Water			Water			Water		
		Saturation	Temperature	Interaction	Saturation	Temperature	Interaction	Saturation	Temperature	Interaction	Saturation	Temperature	Interaction
CO2	F-ratio	2.26**	3.01	1.00	1.87	0.0004	6.73	23.53	1.79	19.52	6.56	0.18	0.01
	p-value	0.15	0.11	0.40	0.20	0.99	0.01*	0.0001*	0.21	0.0002*	0.01*	0.68	0.99
CH_4	F-ratio	29.22	41.63	19.49	17.07	123.21	23.57	1.02	72.03	0.70	0.21	3.42	1.15
	p-value	0.00002*	0.00003*	0.0002*	0.0003*	0.0000001*	0.00007*	0.39	0.000002*	0.51	0.81	0.09	0.35
			1										
					PF (One-way								
		n n	0-10 cm		ANOVA)								
	-	Light/dark	Amendment	Interaction	Amendments								
CO2	F-ratio	16.94	2.87	2.93	1.28								
	p-value	0.003*	0.10	0.10	0.39								
CH_4	F-ratio	4.95	20.95	1.04	0.38								
	p-value	0.06	0.0004*	0.4	0.77								
	915												
	916	* Statistica	ally significan	t values are	marked with a	ι*.							
	917	** All stat	istics show re	esults from tv	vo-way ANO	VA except for	PF, which	shows resu	lts from one-	way ANOV	/A.		

⁹¹³ Table 3. ANOVA results examining microcosms testing water saturation, temperature, light/dark, and amendment effects on CO₂ and

918 Table 4. The cellular oxidation rates normalized to CH₄ concentration from several studies.

		Normalized Cellular Rate
CH₄ (ppmv)	т (°С)	(fmol CH_4 cell ⁻¹ day ⁻¹ ppmv- CH_4^{-1})
\bigcirc	_	This study
1.8	4	0.026
1.8	4	0.005
1.8	4	0.009
2.7	10	0.005
1.8	10	0.056
1.8	10	0.014
1.8	4	0.011
2.7	10	0.004
M		
		Baani and Liesack, 2008
1000	25	0.10
700	25	0.09
600	25	0.08
100	25	0.07
10	25	0.02
1.75	25	0.02
		Knief and Dunfield, 2005
1000	25	0.11
1000	25	0.09
1000	25	0.03
1000	25	0.06
1000	25	0.08



591	28	0.0007
518	21	0.0010
424	12	0.0021
345	4	0.0044
305	0	0.0023

920 Table 5. Extrapolated field fluxes for CO_2 emission and atmospheric CH_4 uptake.

S			((ni	CO_2 Rate mol g ⁻¹ d ⁻¹)	400 4000/ 5-4	Flux mol CO ₂ d^{-1}	400 4000/ 0-1	CH_4 Rate (pmol g ⁻¹ d ⁻¹)				
Reference Rate C	m of depth	kg cryosol	4ºC 100% Sat	10ºC 33% Sat	4ºC 100% Sat	10ºC 33% Sat	4ºC 100% Sat	10ºC 33% Sat	4ºC 100% Sat	10ºC 33% Sat		
0-10 cm	15	189	125.1 ± 13.5	98.2 ± 16.6	0.024±0.002*	0.019±0.003	34 ± 19	260 ± 60	-6.4±3.6*	-49.1±11.3		
30-40 cm	30	486	16.1 ± 2.1	13.0 ± 2.9	0.008±0.001	0.006±0.001	7.5 ± 4.5	42.4 ± 2.7	-3.6±2.2	-20.6±1.3		
60-70 cm	25	360	44.2 ± 2.2	36.5 ± 4.6	0.016±0.001	0.013±0.002	0.7 ± 1.0	7.3 ± 1.8	-0.3±0.4	-2.6±0.6		
70.90 cm	10 144	10 144	10 144	10 144			0.005±0.000	0 002+0 001			0.0+0.1	0 2+1 1
70-80 cm		144	33.5 ± 2.2	20.6 ± 4.6	3	6.0 ± 0.4	0.4 1.6 ± 9.0	-0.9±0.1	-0.2±1.1			
					Total mol	$CO_2 \text{ m}^{-2} \text{ d}^{-1}$			Total µmol	$CH_4 \text{ m}^{-2} \text{ d}^{-1}$		
O)				0.052±0.005	0.041±0.007	-		-11.2±6.2	-72.6±14.4		

921

919

* Assuming a cryosol density of 1.8 g cm⁻³, the total grams of cryosol comprising 1 m⁻² was calculated down to a depth of 80 cm. This mass of cryosol was broken down into sections based on the 4 depths used for determining CO_2 production and CH_4 oxidation rates in the microcosms. Production and consumption was calculated for these 4 zones based on rates obtained from the two end member

- 925 conditions (4°C fully saturated and 10°C unsaturated) and then summed to give a scaled up estimate of total CO₂ emissions and total
- 926 CH₄ oxidation rates. CH₄ total flux assumes penetration of air with a CH₄ concentration above the methanotrophic threshold value.

Author Manuscri





gbi_12193_f2.png





