

1 **The effect of organic carbon on fixed nitrogen loss in the eastern tropical South Pacific and**  
2 **Arabian Sea oxygen deficient zones**

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23 Running head: Nitrogen loss in oxygen deficient zones

24

25 *Acknowledgements*

26 We thank the captains and crews of the R/Vs *Knorr* and *Roger Revelle* for their assistance  
27 collecting the samples for this study. We are grateful to S. W. A. Naqvi for the opportunity to  
28 study the Arabian Sea and to G. Holtgrieve for guiding the statistical analyses. We also  
29 acknowledge our colleagues at the University of Washington Stable Isotope Laboratory for  
30 analytical support. Thoughtful comments from two anonymous reviewers helped improve this  
31 manuscript. Support for this work was provided by National Science Fund grants to AHD and  
32 BBW and the Harry Hess Fellowship of Princeton University to BXC.

33 ***Abstract***

34 The three major oxygen deficient zones (ODZs) of the world oceans (Eastern tropical North  
35 and South Pacific (ETNP and ETSP, respectively), and Arabian Sea (AS) host the vast majority  
36 of pelagic fixed nitrogen (N) loss and up to half of total marine N loss. An important control on  
37 the absolute and relative importance of the two main pathways of N removal (denitrification and  
38 anammox) is thought to be the input of organic matter. In this study, we investigated the  
39 response of N loss in the ETSP and AS ODZs to addition of organic matter in the form of  
40 glucose and naturally-derived dissolved and particulate organic matter (DOM and POM,  
41 respectively). Fresh POM was collected using sediment traps deployed above the ODZ and trap  
42 leachate was used as the DOM treatment. In the ETSP ODZ, the addition of glucose stimulated  
43 denitrification (1.6 fold increase after 5 d) but not anammox (14 fold decrease after 5 d). In the  
44 AS ODZ, only POM, not DOM, significantly increased rates of denitrification at the base of the  
45 oxycline (5.4 – 6.4 fold increase after 2 d), but not at the secondary nitrite maximum. This may  
46 be due to the more refractory nature of the DOM vs. POM (C:N ratio = 9.3 vs. 7.8, respectively)  
47 and a higher carbon consumption rate at the shallower depth. These results suggest that  
48 denitrification was generally limited by organic matter supply at the time of this study in both the  
49 ETSP and AS ODZs, although the lability of the organic matter is important. Interestingly,  $^{15}\text{N}_2$   
50 produced in ETSP and AS incubations were not binomially distributed relative to the reactants  
51 after the influence of anammox was taken into account, suggesting an alternative unknown  
52 production mechanism or pathway of N removal.

53 ***Introduction***

54 The vast majority of pelagic fixed nitrogen (N) removal occurs in three major oxygen  
55 deficient zones (ODZs) of the world: the eastern tropical North and South Pacific (ETNP and  
56 ETSP, respectively), and the Arabian Sea (AS). Although they comprise less than 1% of the  
57 total volume of the ocean (Codispoti et al. 2001), the ODZs are responsible for at least a quarter  
58 of total marine N loss (Codispoti et al. 2001). Wind-driven upwelling stimulates high  
59 productivity in the overlying waters of the ODZs, which sinks and fuels substantial respiration at  
60 depth. Combined with poor ventilation of these regions, the result is a depletion of water column  
61 oxygen to the extent that it becomes thermodynamically favorable for microbes to utilize  $\text{NO}_3^-$ ,  
62  $\text{IO}_3^-$  (Farrenkopf and Luther 2002), oxidized metals (Luther et al. 1997; Moffett et al. 2007), and  
63  $\text{SO}_4^{2-}$  (Canfield et al. 2010) as electron acceptors during organic matter oxidation. Within the  
64 open ocean ODZs, the majority of organic matter respired is ultimately coupled to the reduction  
65 of  $\text{NO}_3^-$ . Although the  $\text{SO}_4^{2-}$  concentration is orders of magnitude higher than the  $\text{NO}_3^-$   
66 concentration, there is no accumulation of  $\text{H}_2\text{S}$ , implying reduced S is efficiently reoxidized,  
67 likely via direct or indirect coupling to the reduction of  $\text{NO}_3^-$  (Canfield et al. 2010). Numerous  
68 transformations of fixed N can occur; however, once converted to  $\text{N}_2\text{O}$  or  $\text{N}_2$ , N becomes  
69 biologically inaccessible except to  $\text{N}_2$  fixing microbes.

70 The two major pathways of fixed N loss in the ODZs are denitrification and the anaerobic  
71 oxidation of ammonium (anammox). Denitrification is a heterotrophic process that proceeds via  
72 a stepwise process in which organic carbon oxidation is coupled to the sequential reduction of N-  
73 oxides to gaseous end products:  $\text{NO}_3^- \rightarrow \text{NO}_2^- \rightarrow \text{NO} \rightarrow \text{N}_2\text{O} (\text{g}) \rightarrow \text{N}_2 (\text{g})$ . Anammox is an  
74 autotrophic process, gaining energy from the oxidation of  $\text{NH}_4^+$  to  $\text{N}_2$  using  $\text{NO}_2^-$ . Due to the  
75 importance of the 3 major ODZs to the balance of fixed N in the ocean, there has been

76 substantial interest in determining the contribution of these two processes to total pelagic N  
77 removal, especially given the potential of these regions to change with changing climate.  
78 <sup>15</sup>N-labeling experiments have suggested that anammox is responsible for the majority of N<sub>2</sub>  
79 production in the ETSP (Thamdrup et al. 2006; Hamersley et al. 2007; Kalvelage et al. 2013) and  
80 AS (Jensen et al. 2011) ODZs. In contrast, other studies in the ETSP (Dalsgaard et al. 2012) and  
81 the AS ODZs (Nicholls et al. 2007; Ward et al. 2009) identified heterotrophic denitrification to  
82 be the dominant N loss pathway. There is evidence that both denitrification and anammox are  
83 regulated by the availability of organic matter in the ODZs (Ward et al. 2008; Kalvelage et al.  
84 2013), and this discrepancy in the relative contributions of denitrification and anammox to N  
85 removal in the ETSP and AS ODZs has been ascribed to the differing responses of anammox and  
86 denitrifying bacteria to organic matter availability (Thamdrup et al. 2006; Ward et al. 2009;  
87 Dalsgaard et al. 2012).

88 Seasonally changing wind patterns in all three ODZs give rise to temporally varying  
89 productivity, which, in turn, affect the timing and quantity of organic matter reaching the ODZs  
90 (Lee et al. 1998). If the flux of organic matter to the ODZs is linked to surface productivity, the  
91 degree to which carbon is available to microbes in either the ETSP or AS ODZ would also be  
92 both spatially and temporally heterogeneous. Denitrifiers have been found to be abundant and  
93 diverse, capable of rapid growth in response to episodic inputs of organic matter (Ward et al.  
94 2008). Anammox bacteria grow more slowly (Van de Graaf et al. 1995) and may maintain a  
95 lower though more constant rate of N removal in the ODZs. Although anammox is itself an  
96 autotrophic process, the substrates anammox depends upon (i.e., NH<sub>4</sub><sup>+</sup> and NO<sub>2</sub><sup>-</sup>) are produced  
97 by the oxidation of organic matter. NH<sub>4</sub><sup>+</sup> is generated at each step of the successive reduction of  
98 N-oxides during heterotrophic denitrification. Additionally, dissimilatory nitrate reduction to

99 ammonium (DNRA, Lam et al. 2009; Jensen et al. 2011) and sulfate reduction (Canfield et al.  
100 2010), both heterotrophic processes, may supply anammox with  $\text{NH}_4^+$ .  $\text{NO}_2^-$  is formed via the  
101 reduction of  $\text{NO}_3^-$  during the oxidation of organic matter.  $\text{NO}_2^-$  might also be produced by  $\text{NH}_4^+$   
102 oxidation (Kalvelage et al. 2013), however this pathway depends on the availability of  $\text{NH}_4^+$  and  
103 thus, organic matter oxidation.

104 The objectives of this work were to investigate the effect of organic matter quality and  
105 availability on the rates of N loss in the ETSP and AS ODZs. In the ETSP ODZ, we measured  
106 the rates and relative contributions of denitrification and anammox in incubations with and  
107 without the addition of a simple organic compound (glucose). In the AS ODZ, we used freshly  
108 collected dissolved organic matter (DOM) and sinking particulate organic matter (POM) from  
109 sediments traps deployed directly above the ODZ in order to assess the response of  $\text{N}_2$   
110 production to naturally derived organic matter. We chose these forms of organic matter because  
111 although bacteria most readily take up DOM, a significant source of DOM in the interior of the  
112 ocean is the degradation of POM. In addition to acting as a source of DOM, bacteria directly  
113 colonize POM and take advantage of living adjacent to both substrate and other bacteria  
114 performing complementary redox transformations (Karl et al. 1984)

115

## 116 ***Methods***

### 117 *Study site and sample collection*

118 Incubation experiments were carried out at one station in the ETSP ODZ aboard the R/V  
119 *Knorr* (October – November 2005) and at two stations in the AS ODZ aboard the R/V *Roger*  
120 *Revelle* (September – October 2007) (Table 1). Samples were collected using a rosette of 10 or

121 30 L Niskin bottles equipped with dissolved oxygen ( $O_2$ ) sensors calibrated by Winkler  
122 titrations, in addition to conductivity, temperature, and pressure sensors.

123 Sampling depths were chosen based on  $O_2$  and nitrite ( $NO_2^-$ ) concentrations: at the secondary  
124  $NO_2^-$  maximum (SNM), and at the shallowest depth where  $O_2$  was undetectable (base of the  
125 oxycline; in AS only). The base of the oxycline was chosen based on previous studies, which  
126 have reported the highest rates of  $N_2$  production at the top of the ODZ, possibly due to the  
127 relatively high flux of POM compared to deeper in the ODZ. The SNM was also sampled as it is  
128 associated with the most oxygen depleted waters (Thamdrup et al. 2012) and has classically been  
129 regarded as a zone of active N loss (Codispoti and Christensen 1985).

130 Incubations were carried out in duplicate in acid-washed, large-volume, gas-tight trilaminate  
131 bags (Pollution Management Corporation) fitted with three-way stopcocks. Prior to sampling,  
132 bags were flushed at least three times with  $CO_2$  to eliminate  $O_2$  from any possible headspace and  
133 evacuated. Approximately 8 L of water were gravity-fed into each bag directly from a Niskin  
134 bottle, taking care to prevent any contact with the atmosphere. The headspace of each Niskin  
135 bottle was continuously flushed with  $CO_2$  through the vent to prevent atmospheric  $O_2$   
136 contamination during sampling. Bags received additions of  $^{15}N$ -labeled  $NO_3^-$  (in ETSP only),  
137  $NO_2^-$  (in AS only) or  $NH_4^+$  (in both, Table 2).

138 In the ETSP, dissolved organic matter (DOM) in the form of glucose was also added to  
139  $^{15}NO_3^-$  and  $^{15}NH_4^+$  labeled incubations to a final concentration of  $2 \mu\text{mol C L}^{-1}$ . In the AS  
140 experiments,  $^{15}NO_2^-$  labeled incubations were amended with POM and DOM collected in situ  
141 from the AS ODZ (*see* below for details of collection and preparation) with and without  $^{14}NH_4^+$ .  
142 Tracers and amendments were added while filling each bag either by injecting directly into the  
143 Tygon tubing connecting the Niskin to the bag or through the three-way valve. POM was added

144 to the bags to a final concentration of  $2.6 \mu\text{mol C L}^{-1}$  and  $0.33 \mu\text{mol N L}^{-1}$  at Sta. 1 and  $1.3 \mu\text{mol}$   
145  $\text{C L}^{-1}$  and  $0.17 \mu\text{mol N L}^{-1}$  at Sta. 2, determined by CHN elemental analyzer. At both Sta. 1 and  
146 2, DOM was added to the bags to a final concentration of  $0.24 \mu\text{mol C L}^{-1}$  and  $0.026 \mu\text{mol N L}^{-1}$ ,  
147 respectively, determined by total organic carbon and nitrogen analyzer (TOC-V/Total Nitrogen  
148 M-1 Shimadzu).

149 For the AS experiments, POM was collected at a depth in the oxycline, approximately 20 and  
150 50 m above the ODZ proper at Sta. 1 and 2, respectively, using a NetTrap (Petersen et al. 2005),  
151 a large diameter ( $\sim 2$  m), free-floating sediment trap, based on the design of a closing plankton  
152 net, capable of collecting large amounts of sinking POM ( $> 50 \mu\text{m}$ ) in relatively short time  
153 periods (24 – 36 h). Contents of the cod end were filtered through a glass fiber filter ( $0.7 \mu\text{m}$   
154 nominal pore size). Filtrate (trap leachate) was collected for use as the DOM amendment. The  
155 material on the filter was resuspended into a smaller quantity of site water for use as the POM  
156 amendment. Each organic matter addition was degassed before introduction to the bag by  
157 applying a vacuum to cause continuous boiling for  $\sim 15$  min.

158 Ambient and tracer concentrations of  $\text{NO}_3^-$ ,  $\text{NO}_2^-$ , and  $\text{NH}_4^+$  in all incubations were measured  
159 by autoanalyzer using standard colorimetric techniques (Strickland and Parsons 1972).

160 Additionally, in the ETSP incubations,  $\text{NO}_3^-$  and  $\text{NO}_2^-$  were measured by autoanalyzer at  
161 approximately 12 h intervals for the length of the incubation. Detection limits of  $\text{NO}_3^-$ ,  $\text{NO}_2^-$ ,  
162 and  $\text{NH}_4^+$  were  $0.08$ ,  $0.01$ , and  $0.07 \mu\text{mol L}^{-1}$ , respectively.

163

#### 164 *<sup>15</sup>N-labeled incubations*

165 Samples for the determination of  $^{29}\text{N}_2$  and  $^{30}\text{N}_2$  were taken following the method of Emerson  
166 et al. (1999). Approximately 150 mL of water was drawn from the Niskin bottles (representing



167 initial N<sub>2</sub> gas and isotope concentrations) or from a bag into 300 mL HgCl<sub>2</sub>-poisoned, pre-  
168 evacuated glass flasks equipped with a 9 mm, gas-tight, single o-ring valves (Louwers-Hapert).  
169 The flasks were returned to the University of Washington where they were weighed and  
170 dissolved gases were equilibrated with the headspace of the flask at a constant temperature for 24  
171 hours. The headspace gases were transferred to a stainless steel finger immersed in liquid He,  
172 during which water and CO<sub>2</sub> were trapped cryogenically. Samples were analyzed on a Finnigan  
173 Delta XL dual inlet isotope ratio mass spectrometer for mass ratios 29:28, 30:28, 28:40, relative  
174 to an in-house gas standard with known gas and isotope ratios. Anammox and denitrification  
175 rates were calculated from the production of <sup>15</sup>N-labeled N<sub>2</sub> following de Brabandere et al.  
176 (2013).

177

#### 178 *Statistical analyses*

179 The effects of treatments on the rates of N<sub>2</sub> production in the AS incubations were examined  
180 using an analysis of variance (ANOVA). The factors affecting N<sub>2</sub> production that were  
181 considered were location (Sta. 1 or 2), depth (base of oxycline or SNM), <sup>14</sup>NH<sub>4</sub><sup>+</sup> (with or  
182 without), and organic matter (control, POM, and DOM). A post hoc comparison using Tukey's  
183 HSD test was used to evaluate differences in organic matter treatments. All statistical analyses  
184 were performed using SPSS version 13.

185

#### 186 **Results**

##### 187 *Hydrographic conditions*

188 At the ETSP study site, the ODZ (defined as the minimum value reached by the Seabird  
189 oxygen sensor mounted to the sampling rosette) extended from approximately 75 to 400 m. The

190 mixed layer was shallow ( $< 20$  m, defined as a change in  $\sigma_t > 0.03$  kg m<sup>-3</sup> [de Boyer Montégut et  
191 al. 2004]), sea surface temperature (SST) was relatively low (15.3°C), and [NO<sub>3</sub><sup>-</sup>] and [PO<sub>4</sub><sup>3-</sup>]  
192 were relatively high in surface waters (5.9 and 0.8  $\mu\text{mol L}^{-1}$ , respectively), all of which are  
193 indicative of active upwelling. [NO<sub>2</sub><sup>-</sup>] was detectable at the surface (0.2  $\mu\text{mol L}^{-1}$ ) and formed a  
194 broad secondary maximum within the ODZ (maximum 4.4  $\mu\text{mol L}^{-1}$ ). A distinct primary NO<sub>2</sub><sup>-</sup>  
195 maximum (PNM) was not detected, probably due to the sampling distribution being too coarse to  
196 resolve the narrow peak. [NH<sub>4</sub><sup>+</sup>] was as high as 0.3  $\mu\text{mol L}^{-1}$  in the mixed layer, decreasing to  
197 below detection ( $< 0.05$   $\mu\text{mol L}^{-1}$ ) in and below the ODZ.

198 The AS field campaign took place at the end of the high productivity southwest monsoon,  
199 although conditions were already transitioning to the more oligotrophic fall intermonsoon as  
200 indicated by relatively high SST (28.8°C) and a shallow mixed layer (40 m) in which [NO<sub>3</sub><sup>-</sup>] and  
201 [PO<sub>4</sub><sup>3-</sup>] were undetectable. Hydrographic conditions were previously reported by Ward et al.  
202 (2009), which we will summarize here. A distinct PNM was present at the base of the mixed  
203 layer (maximum 0.9 and 2.8  $\mu\text{mol L}^{-1}$ , at Sta. 1 and 2, respectively), underlain by a much larger  
204 SNM (maximum 5.7 and 7.7  $\mu\text{mol L}^{-1}$ , at Sta. 1 and 2, respectively), which extended from  
205 approximately 100-150 to 400 m. At the more northern station sampled (Sta. 1), O<sub>2</sub> was  
206 undetectable from approximately 100 to 800 m. At the more southern station sampled (Sta. 2),  
207 the ODZ was slightly thinner, extending from 150 to 800 m. [NH<sub>4</sub><sup>+</sup>] was as high as 0.6  $\mu\text{mol L}^{-1}$   
208 in the mixed layer, decreasing to below detection ( $< 0.05$   $\mu\text{mol L}^{-1}$ ) in and below the ODZ.

209

#### 210 *N transformations and loss in ETSP ODZ incubations*

211 At Sta. 20 in the ETSP, in the incubations with glucose, the initial concentration of 24  $\mu\text{mol}$   
212 L<sup>-1</sup> NO<sub>3</sub><sup>-</sup> was rapidly drawn down to  $< 1$   $\mu\text{mol L}^{-1}$  in less than 3 days, and undetectable by 5.5

213 days (Fig. 1a). There was some  $\text{NO}_2^-$  accumulation until 1.5 days, after which  $\text{NO}_2^-$  was quickly  
214 consumed to undetectable levels by 4.5 days (Fig. 1b). This pattern of nutrient consumption was  
215 in contrast to the incubation in which no organic carbon was added. In these incubations,  $\text{NO}_3^-$   
216 remained unchanged or even slightly increased until approximately 3 days, after which  $\text{NO}_3^-$   
217 steadily decreased to undetectable levels by 11.5 days. Nitrite decreased  $\sim 1 \mu\text{mol L}^{-1}$  for one day  
218 before slowly increasing to maximum values by 9.5 – 10 days followed by rapid consumption to  
219 undetectable levels by 11.5 days.

220 Samples for  $^{15}\text{N}_2$  production were taken at 5.7 days (Fig. 2). In the control incubations (only  
221  $^{15}\text{N}$  tracers added), the average denitrification rate over the length of the incubation was  $2.6 \text{ nmol}$   
222  $\text{N}_2 \text{ h}^{-1}$ , and the average anammox rate was  $0.9 \text{ nmol N}_2 \text{ h}^{-1}$  over the course of the incubation. In  
223 the incubations amended with glucose, the average denitrification rate was  $4.2 \text{ nmol N}_2 \text{ h}^{-1}$   
224 during the length of the incubation and the average anammox rate was  $0.06 \text{ nmol N}_2 \text{ h}^{-1}$  over the  
225 course of the incubation.

226

#### 227 *Sediment trap organic matter composition in the AS*

228 Arabian Sea DOM (trap leachates) had an atomic C:N ratio of 9.3. This ratio is higher than  
229 that of the particulate trap flux, which had an average C:N ratio of 7.8.

230

#### 231 *N loss in AS ODZ incubations*

232 The addition of  $^{14}\text{NH}_4^+$  had no significant effect on  $^{29}\text{N}_2$  or  $^{30}\text{N}_2$  production in the AS  
233 regardless of station, depth, or organic matter addition ( $F_{1,21} = 0.486$ ,  $p > 0.05$ , data not shown).  
234 Since the addition of  $^{14}\text{NH}_4^+$  had no significant effect on  $^{15}\text{N}_2$  production rates, all rates with and

235 without organic matter additions are presented as an average of the treatments with and without  
236  $^{14}\text{NH}_4^+$ .

237 In the AS incubations, samples for  $^{15}\text{N}_2$  production were taken at approximately 2 days.  $^{15}\text{N}_2$   
238 production rates by anammox (in  $^{15}\text{NH}_4^+$  amended incubations) and denitrification (in  $^{15}\text{NO}_2^-$   
239 amended incubations with no organic matter additions) were previously published in Ward et al.  
240 (2009). We present them here as the ‘control’ to assess the effect of organic matter additions on  
241  $\text{N}_2$  production rates. At Sta. 1, in the control incubations (no organic matter additions), the  
242 average denitrification rates over the length of the incubations were  $0.21 \pm 0.14$  and  $0.15 \pm 0.06$   
243  $\text{nmol N}_2 \text{ h}^{-1}$ , at the base of the oxycline and SNM, respectively (Fig. 3a). At Sta. 2, in the control  
244 incubations, the average denitrification rates over the course of the incubations were  $0.06 \pm 0.02$   
245 and  $0.05 \pm 0.02 \text{ nmol N}_2 \text{ h}^{-1}$ , at the base of the oxycline and SNM, respectively (Fig. 3b).

246 The two organic carbon treatments, DOM and POM collected freshly from sediment traps,  
247 had different effects on denitrification rates (determined from  $^{30}\text{N}_2$  production) in the AS ODZ  
248 (Fig. 3a,b). Tukey’s post hoc test revealed that denitrification rates with and without DOM were  
249 not significantly different ( $p > 0.05$ ). The addition of POM from sediment traps significantly  
250 increased denitrification rates at both stations relative to the control ( $p < 0.05$ ) at the base of the  
251 oxycline only ( $F_{1,10} = 10.255, p < 0.01$ ), however the response of denitrification to the addition  
252 of POM was not significantly different between the two stations ( $F_{1,10} = 0.565, p > 0.05$ ).  
253 Average denitrification rates in the POM treatment at the base of the oxycline was 6.4 times  
254 larger than the control at Sta. 1 and 5.4 times larger at Sta. 2.

255 Parallel incubations with  $^{15}\text{NH}_4^+$  only without any additions of organic matter were carried  
256 out at each station and depth in the AS (Fig. 3a,b), which allowed the determination of the  
257 control anammox rate. At Sta. 1, the average anammox rates during incubations were  $0.005 \pm$

258 0.0004 and  $0.007 \pm 0.0009$  nmol N<sub>2</sub> h<sup>-1</sup>, at the base of the oxycline and SNM, respectively. At  
259 Sta. 2, the average anammox rates during incubations were  $0.009 \pm 0.008$  and  $0.06 \pm 0.05$  nmol  
260 N<sub>2</sub> h<sup>-1</sup>, at the base of the oxycline and SNM, respectively.

261

## 262 ***Discussion***

### 263 *Effect of dissolved organic carbon (glucose) on N loss in the ETSP*

264 Concentrations of DOC in the Pacific Ocean range from as high as 70 μmol L<sup>-1</sup> in the low  
265 latitude surface ocean to < 40 μmol L<sup>-1</sup> in the deep ocean (Hansell and Carlson 1998). The  
266 amount of glucose added to incubations in this study (2 μmol C L<sup>-1</sup>) was only a small fraction of  
267 the DOC naturally present. However, this ambient DOC is thought to be very old (4000 – 6000  
268 y B. P.), estimated from radiocarbon measurements (Druffel et al. 1992), indicating that the  
269 majority of this DOC is refractory and is returned to the deep unrespired after a complete ocean  
270 mixing cycle. Glucose (C<sub>6</sub>H<sub>12</sub>O<sub>6</sub>) is a simple carbohydrate that is a readily utilizable energy  
271 source by microorganisms that yields ATP via glycolysis, thus, even a relatively small amount of  
272 glucose may fuel relatively high rates of bacterial respiration.

273 At Sta. 20 in the ETSP, both the control and glucose-amended incubations exhibited NO<sub>3</sub><sup>-</sup>  
274 consumption, and NO<sub>2</sub><sup>-</sup> accumulation and consumption in a classic denitrifying sequence, as  
275 observed in bacterial cultures, which is controlled by the enzyme kinetics of each step of  
276 denitrification (Betlach and Tiedje 1981). The addition of glucose to these incubations  
277 stimulated NO<sub>3</sub><sup>-</sup> and NO<sub>2</sub><sup>-</sup> reduction by supplying substrate to a carbon-limited system (Fig.  
278 1a,b). The average denitrification rate measured by the production of <sup>15</sup>N<sub>2</sub> was 1.6 times higher  
279 in the incubations with the addition of glucose compared to without (Fig. 2). Conversely, the  
280 average anammox rate was an order of magnitude lower in the glucose treatments compared to

281 the control. Although glucose does not contain organic N, which would be remineralized to  
282  $\text{NH}_4^+$  potentially for anammox, it is unlikely that anammox bacteria in these incubations were  
283  $\text{NH}_4^+$ -limited due to the addition of  $1 \mu\text{mol L}^{-1} \text{ }^{15}\text{NH}_4^+$  tracer.

284 These results support the findings of Dalsgaard et al. (2012) in the ETSP ODZ, who  
285 hypothesize a lack of close coupling between denitrification and anammox based on the  
286 observation that denitrification rates were high when anammox rates were low and vice versa.  
287 Dalsgaard et al. (2012) argue that the lack of close coupling may be due to the relative response  
288 times of the different microorganisms involved: denitrifiers are fast-growing and can respond  
289 quickly to episodic inputs of organic matter, whereas anammox bacteria are relatively slow-  
290 growing which limits the rate of their response to increased substrate; however, they  
291 subsequently maintain lower rates for longer periods of time.

292

#### 293 *Effect of $\text{NH}_4^+$ on denitrification in the AS*

294 Previous researchers have also found no significant increase in the anammox rate (determined  
295 from  $^{29}\text{N}_2$  production) in  $^{15}\text{NO}_{2,3}^-$  labeled incubations with  $^{14}\text{NH}_4^+$  compared to without (Kuypers  
296 et al. 2005; Thamdrup et al. 2006). This is somewhat unexpected since  $[\text{NH}_4^+]$  in both the  
297 Benguela upwelling and ETSP ODZs is frequently below detection suggesting anammox could  
298 be limited by  $[\text{NH}_4^+]$ . Thamdrup et al. (2006) speculated that no stimulation of anammox by the  
299 addition of  $\text{NH}_4^+$  may reflect extremely efficient ammonium uptake by anammox bacteria, which  
300 would lead to reaching saturating concentrations at a relatively low  $[\text{NH}_4^+]$ .

301

#### 302 *Effect of dissolved organic matter on denitrification in the AS*

303 The absence of a significant response from the addition of DOM indicated that denitrifiers in  
304 the AS ODZ were either restricted in their ability to respire the DOM added due to its possibly  
305 refractory nature (Druffel et al. 1992), or did not receive sufficient additional DOM to cause a  
306 response. Due to the significant increase in denitrification rates by the addition of POM (*see*  
307 discussion in following section), it is unlikely that denitrifiers had no response to the addition of  
308 DOM because they were not carbon-limited.

309 The C:N ratio of the DOM was high (9.3) compared to both the Redfield ratio (6.6) and the  
310 POM collected concurrently in the sediment trap (7.8), suggesting the DOM was more  
311 chemically degraded. Labile N-rich proteins and amino acids are preferentially remineralized  
312 leaving behind the more refractory fraction of DOM (Walker and McCarthy 2012). Another  
313 possibility is that not enough DOM was added to cause a response. Approximately an order of  
314 magnitude less organic C and N were added to DOM amended incubations compared to POM  
315 treatments. Although the aim was to add a comparable amount of organic matter to all POM and  
316 DOM treatments, the exact composition and concentration of both the POM and DOM were  
317 determined only upon return to a land-based laboratory and as such, was unknown at the time the  
318 incubations were being carried out.

319

#### 320 *Effect of particulate organic matter on denitrification in the AS*

321 Stimulation by the addition of only the particulate form of organic matter suggests that we  
322 may have essentially inoculated the bag incubations with denitrifying microbes associated with  
323 the POM, or provided ample labile organic carbon for respiration, or both. Previous work  
324 analyzing the chemical composition of sinking POM captured in sediment traps has concluded  
325 that sinking POM is composed of more labile organic matter relative to suspended POM or

326 DOM. A significant increase in denitrification rates was observed at the base of the oxycline  
327 only, and not at the SNM, at both stations suggesting that it was the POM itself, not particle  
328 associated denitrifiers that gave rise to increased denitrification rates. If the stimulation were due  
329 to new bacteria introduced with the POM, a comparable absolute increase in rates at both depths  
330 at each station would be expected. Additionally, although double the amount of POM was added  
331 to incubations at Sta. 1 relative to Sta. 2, the proportional increase in the average denitrification  
332 rates relative to the control at the base of the oxycline of both stations was similar (6.4 vs. 5.4  
333 fold increase at Sta. 1 and 2, respectively) suggesting the original bacterial population in each  
334 incubation had comparable responses to the addition of POM. However, the absolute increase in  
335 denitrification rates at Sta. 1 was triple that of Sta. 2 ( $1.1$  vs.  $0.3 \text{ nmol L}^{-1} \text{ h}^{-1}$ , respectively),  
336 although double the POM was added at Sta. 1 relative to Sta. 2, further suggesting that the  
337 increased denitrification rates were due to a stimulation of the original bacterial population in the  
338 bags, not due to bacteria introduced with the POM.

339 The flux of organic matter is correlated to the rate of fixed N removal in the ETSP and AS  
340 ODZs (Jensen et al. 2011; Kalvelage et al. 2013) suggesting organic matter is an important  
341 control on this process. Further supporting this conclusion is the observation that rates of  $\text{N}_2$   
342 production generally decrease with increasing depth in both these regions (Thamdrup et al. 2006;  
343 Jensen et al. 2011; Dalsgaard et al. 2012), following the trend of decreasing organic matter flux  
344 with increasing depth (Martin et al. 1987). Additionally, carbon limitation of denitrification in  
345 the ETSP has been directly measured in incubations (Ward et al. 2008). In the present study, we  
346 hypothesize that the significant stimulation of denitrification at the base of the oxycline only, not  
347 the SNM, by the addition of POM is due to increased respiration leading to elevated organic  
348 carbon demand at the shallower depth.



349 The magnitude of the POM additions ( $2.6$  and  $1.3 \mu\text{mol C L}^{-1}$  at both depths at Sta. 1 and 2,  
350 respectively) can be evaluated in the context of ambient POC fluxes and remineralization rates in  
351 the AS. The Martin equation (Martin et al. 1987):

$$352 \quad \text{POC flux at depth } z = \text{POC flux at } 100\text{m} \times (z/100)^{-b} \quad (1)$$

353 was used to estimate the POC flux to the depths sampled in this study. Temporal patterns of  
354 productivity and POC export in the AS are dominated by seasonal monsoonal cycles with the  
355 highest productivity and export during the Northeast and Southwest monsoons (Lee et al. 1998).  
356 A  $^{234}\text{Th}$ -based average export from a comparable depth, location, and season as this study (100  
357 m at an open ocean AS ODZ station during the SW monsoon) is  $10.85 \text{ mmol C m}^{-2} \text{ d}^{-1}$  (Lee et al.  
358 1998). An estimate of the attenuation coefficient ( $b$ ) in Eq. 1 at the same open ocean AS ODZ  
359 station is  $0.74$  (Berelson 2001). In order to specify a depth interval over which POC is  
360 consumed, the 8 L bag incubation is assumed to be a cube with a height of  $0.2$  m. Using these  
361 values, the average in situ rates of POC consumption in these incubations would be  $0.08$ ,  $0.04$ ,  
362 and  $0.02 \mu\text{mol C L}^{-1} \text{ d}^{-1}$  at  $100$ ,  $150$ , and  $200$  m. These calculated POC consumption rates are  
363 comparable to measured bacterial carbon demand associated with different forms of POC (Smith  
364 et al. 1992). Although the POC additions in this study exceeded the estimated POC consumption  
365 rate at all depths, it may be that the base of the oxycline harbors a more active and/or denser  
366 bacterial population relative to the deeper SNM due to a chronically larger POC flux and thus,  
367 were able to produce a larger response to the addition of fresh POC.

368

### 369 *Amammox and the source of excess $^{29}\text{N}_2$ production in ETSP and AS incubations*

370 During the 2005 R/V *Knorr* cruise to the ETSP, additional  $^{15}\text{NO}_3^-$  amended bag incubations  
371 were carried out at other stations using the same method described here (data not shown). After

372 2 days,  $^{29}\text{N}_2$  and  $^{30}\text{N}_2$  produced by denitrification were binomially distributed after taking the  
373 production due to anammox into account, such that the total  $^{29}\text{N}_2$  produced was equal to the sum  
374 of  $^{29}\text{N}_2$  from anammox (determined from a parallel incubation with  $^{15}\text{NH}_4^+$ ) and  $^{29}\text{N}_2$  from  
375 denitrification (predicted from  $^{30}\text{N}_2$  assuming a binomial distribution of  $\text{N}_2$  relative to the initial  
376 fraction labeled of  $^{15}\text{NO}_3^-$ ).

377 However, this was not true after 5.7 days in the ETSP incubations (Fig. 2) and in all of the AS  
378 incubations after 2 days (Fig. 4a,b). In these incubations more  $^{29}\text{N}_2$  was produced than could be  
379 accounted for based on the fraction of the initial  $\text{NO}_{2,3}^-$  that was labeled, after taking into account  
380 the production due to anammox. We name this excess the residual  $^{29}\text{N}_2$  ( $\Delta_{\text{resid}}$ ). It is defined as  
381 the difference between the total measured  $^{29}\text{N}_2$  production rate and the  $^{29}\text{N}_2$  production rate  
382 predicted from the sum of anammox and denitrification rates, assuming that the isotopic  
383 composition of the products is binomially distributed relative to the reactants.

384 Nicholls et al. (2007) in the AS found the isotopic composition of  $^{15}\text{N}_2\text{O}$  produced by  
385 denitrification was binomially distributed relative to the starting pool of  $\text{NO}_2^-$ . However, similar  
386 to the results in this study,  $^{15}\text{N}_2$  production could not be predicted by the binomial distribution.  
387 Trimmer and Purdy (2012) measured  $\text{N}_2$  production that was not via canonical denitrification or  
388 anammox, and hypothesized amine groups on allylthiourea (ATU) were directly oxidized to  $\text{N}_2$   
389 by  $\text{NO}_2^-$ . In incubations with  $^{15}\text{NO}_2^-$  in the ETSP, de Brabandere et al. (2013) observed  $^{15}\text{N}_2$   
390 production that was non-binomially distributed after production by anammox had been taken into  
391 account. These researchers speculated that production of  $^{29}\text{N}_2$  via denitrification in the  $^{15}\text{NO}_2^-$   
392 incubations might be underestimated if denitrifiers reduced ambient  $\text{NO}_3^-$  directly to  $\text{N}_2$   
393 intracellularly, without allowing the  $\text{NO}_2^-$  to mix completely with the ambient  $\text{NO}_2^-$  pool ('nitrite  
394 shunting').

395 Production of  $^{29}\text{N}_2$  in excess of the predicted binomial distribution indicates that the  $^{15}\text{N}$ -  
396 labeled fraction of the reactant pool is not well known. To further explore possible mechanisms  
397 responsible for the production of  $\Delta_{\text{resid}}$ , we calculated the average  $^{15}\text{N}$ -labeled fraction of the  
398 reactant pool that would have been required to produce the observed  $^{29}\text{N}_2$  and  $^{30}\text{N}_2$ , with no  $\Delta_{\text{resid}}$ ,  
399 assuming the anammox rates measured in the control incubations remained constant across all  
400 treatments (this is reasonable considering the slow growth rate of anammox bacteria relative  
401 denitrifiers) and  $^{15}\text{N}_2$  was binomially distributed relative to the reactants. In the ETSP  
402 incubations, the fraction of the  $\text{NO}_3^-$  pool that was  $^{15}\text{N}$ -labeled was approximately 0.1. In order  
403 to produce the observed distributions of  $^{29}\text{N}_2$  and  $^{30}\text{N}_2$  at 5.7 d, the  $^{15}\text{N}$ -labeled fraction of the  
404 reactant pool must have been much lower: 0.002. In the AS incubations, the fraction of the  $\text{NO}_2^-$   
405 pool  $^{15}\text{N}$ -labeled was between 0.4 – 0.7. For  $\Delta_{\text{resid}} = 0$ , the fraction of the reactant pool  $^{15}\text{N}$ -  
406 labeled must have been 0.001 – 0.5. These results indicate that direct oxidation of organic N to  
407  $\text{N}_2$  is probably not the only mechanism producing the observed distributions of  $^{15}\text{N}_2$ , given that  
408 an organic  $^{14}\text{N}$  pool up to 1000 times greater than the  $^{15}\text{NO}_{2,3}^-$  concentration would be required.  
409 Thus in this study, ‘nitrite shunting’ (de Brabandere et al. 2013) is a reasonable mechanism to  
410 explain a significant portion of the  $\Delta_{\text{resid}}$ . At 2 days in the ETSP incubations,  $^{29}\text{N}_2$  and  $^{30}\text{N}_2$  were  
411 binomially distributed, which is consistent with ‘nitrite shunting’ not affecting the predicted  
412 distribution of  $^{29}\text{N}_2$  and  $^{30}\text{N}_2$  since  $^{15}\text{NO}_3^-$  was used as the tracer. At 5.7 d,  $\Delta_{\text{resid}}$  becomes  
413 significant which may be due to changes in the fraction  $^{15}\text{N}$ -labeled of the substrates, suggested  
414 by significant changes in the concentrations of  $\text{NO}_3^-$  and  $\text{NO}_2^-$ . Additionally, similar to  
415 denitrification rates,  $\Delta_{\text{resid}}$  was significantly affected by only the addition of POM ( $p < 0.005$ ); no  
416 significant effect was observed between stations, depths, or with the addition of  $\text{NH}_4^+$ . This  
417 response suggests a heterotrophic source for  $\Delta_{\text{resid}}$ .

418

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520



**Table 1**

Station locations, sampling depths, and associated hydrographic characteristics.  $[O_2]$  measured by Seabird  $O_2$  sensor, detection limit  $\sim 1 \mu\text{mol L}^{-1}$ . Detection limits of  $\text{NO}_3^-$ ,  $\text{NO}_2^-$ , and  $\text{NH}_4^+$  were 0.08, 0.01, and  $0.07 \mu\text{mol L}^{-1}$ , respectively.

Station	Latitude	Longitude	Bottom depth (m)	Sampling depth (m)	Feature	$[O_2]$ ( $\mu\text{mol L}^{-1}$ )	$[\text{NO}_3^-]$ ( $\mu\text{mol L}^{-1}$ )	$[\text{NO}_2^-]$ ( $\mu\text{mol L}^{-1}$ )	$[\text{NH}_4^+]$ ( $\mu\text{mol L}^{-1}$ )
ETSP									
20	13.3°S	77°W	788	260	SNM	nd	24.3	4.4	0.1
AS									
1	19.4°N	66.7°E	3095	100	base of oxycline	nd	22.3	1.9	nd
				150	SNM	nd	15.1	5.7	0.06
2	15°N	64°E	3900	150	base of oxycline	nd	10.7	3.7	nd
				200	SNM	nd	7.6	7.7	nd

nd – not detected

521  
522

**Table 2**Summary of  $^{15}\text{N}$ -labeled tracer additions and treatments.

Station	Feature	$^{15}\text{N}$ -tracer <sup>a</sup>	Treatment
ETSP			
20	SNM	$^{15}\text{NO}_3^-$ (2)	–
		$^{15}\text{NO}_3^-$ (2)	glucose
		$^{15}\text{NH}_4^+$ (1)	–
		$^{15}\text{NH}_4^+$ (1)	glucose
AS			
1	SNM, base of oxycline	$^{15}\text{NO}_2^-$ (5)	–
		$^{15}\text{NO}_2^-$ (5)	$^{14}\text{NH}_4^+$
		$^{15}\text{NO}_2^-$ (5)	POM <sup>b</sup>
		$^{15}\text{NO}_2^-$ (5)	POM + $^{14}\text{NH}_4^+$ (5)
		$^{15}\text{NO}_2^-$ (5)	DOM <sup>c</sup>
		$^{15}\text{NO}_2^-$ (5)	POM + $^{14}\text{NH}_4^+$ (5)
		$^{15}\text{NH}_4^+$ (5)	$^{14}\text{NO}_2^-$
2	SNM, base of oxycline	$^{15}\text{NO}_2^-$ (5)	–
		$^{15}\text{NO}_2^-$ (5)	$^{14}\text{NH}_4^+$
		$^{15}\text{NO}_2^-$ (5)	POM
		$^{15}\text{NO}_2^-$ (5)	POM + $^{14}\text{NH}_4^+$ (5)
		$^{15}\text{NO}_2^-$ (5)	DOM
		$^{15}\text{NO}_2^-$ (5)	POM + $^{14}\text{NH}_4^+$ (5)
		$^{15}\text{NH}_4^+$ (5)	$^{14}\text{NO}_2^-$

<sup>a</sup> Target concentration of  $^{15}\text{N}$  tracer and  $^{14}\text{NH}_4^+$  in parentheses ( $\mu\text{mol L}^{-1}$ ); <sup>b</sup> POM – particulate organic carbon; <sup>c</sup> DOM – dissolved organic carbon. Both organic matter treatments came from sediment traps deployed approximately 20 and 50 m above the ODZ proper at Sta. 1 and 2, respectively. *See text for details of preparation.*

523 **Figure legends**

524

525 Figure 1: (A)  $[\text{NO}_3^-]$  and (B)  $[\text{NO}_2^-]$  in incubations with and without glucose added at Sta. 20 in  
526 the ETSP ODZ. Vertical dashed line indicates when sample for  $^{15}\text{N}_2$  was taken. Error bars  
527 represent range of duplicates.

528

529 Figure 2: Denitrification, anammox, and  $\Delta_{\text{resid}}$  rates in incubations with and without glucose  
530 added at Sta. 20 in the ETSP ODZ.

531

532 Figure 3: Denitrification rates in control, DOM, and POM amended incubations at (A) Sta. 1 and  
533 (B) Sta. 2 in the AS ODZ. Also shown are anammox rates measured in parallel incubations with  
534  $^{15}\text{NH}_4^+$ . Error bars are 1 standard deviation (SD). Anammox and denitrification (control only)  
535 rates previously published in Ward et al. (2009).

536

537 Figure 4:  $\Delta_{\text{resid}}$  in control, DOM, and POM amended incubations at (A) Sta. 1 and (B) Sta. 2 in  
538 the AS ODZ.  $\Delta_{\text{resid}}$  is the  $^{29}\text{N}_2$  production rate in excess of the amount predicted from the  $^{30}\text{N}_2$   
539 production rate in  $^{15}\text{NO}_2^- + ^{14}\text{NH}_4^+$  incubations, assuming the  $\text{N}_2$  isotopomers generated by  
540 denitrification are binomially distributed.  $^{29}\text{N}_2$  production from anammox was taken into  
541 account in the calculation of  $\Delta_{\text{resid}}$ . Error bars are 1 SD.

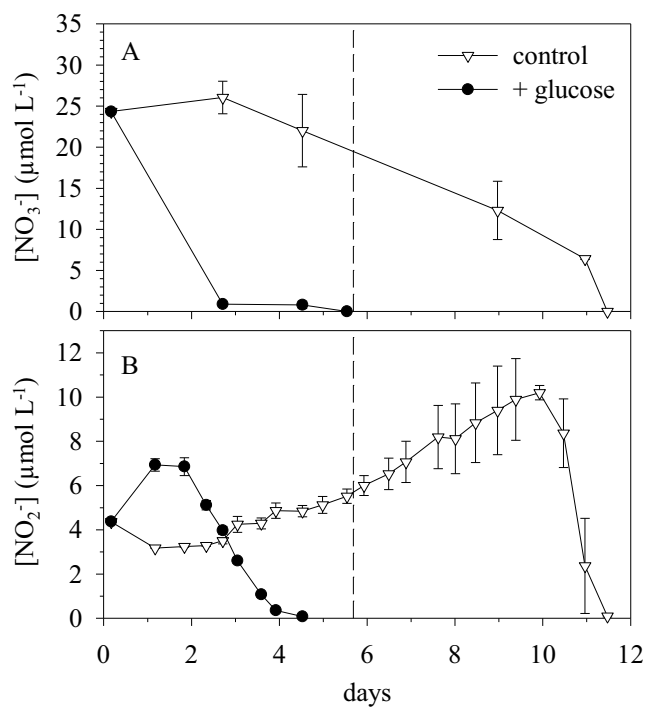


Figure 1

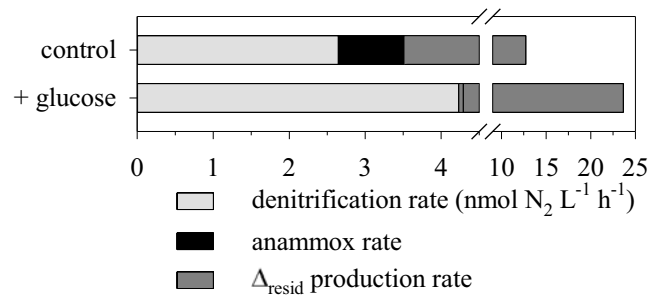


Figure 2

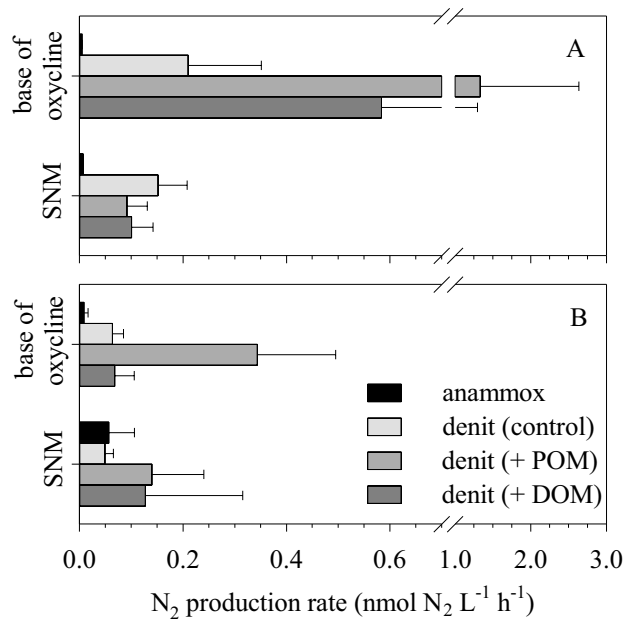


Figure 3

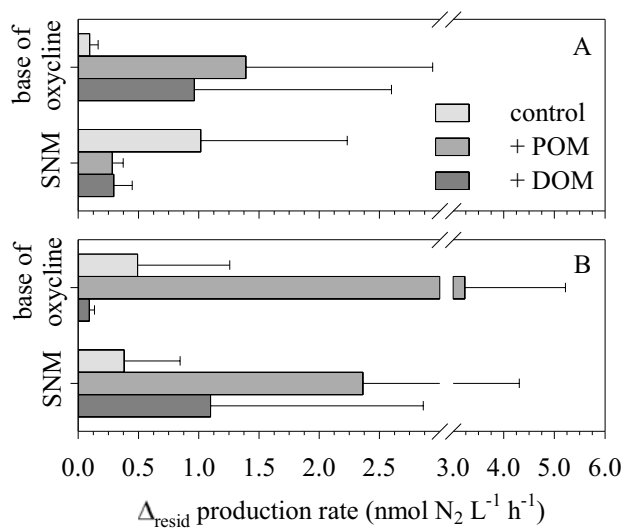


Figure 4